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1 The cichlid-*Cichlidogyrus* network: A blueprint for a model system of parasite evolution

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12

13 **Abstract**

14 Species interactions are a key aspect of evolutionary biology. Parasites, specifically, are drivers of the evolution
15 of species communities, and impact biosecurity and public health. However, when using interaction networks for
16 evolutionary studies, interdependencies between distantly related species in these networks are shaped by ancient
17 and complex processes. We propose using recent interacting host-parasite radiations, e.g. African cichlid fishes
18 and cichlid gill parasites belonging to *Cichlidogyrus* (Dactylogyridae, Monogenea), as macroevolutionary model
19 of species interactions. The cichlid-*Cichlidogyrus* network encompasses 138 parasite species and 416 interactions
20 identified through morphological characteristics and genetic markers in 160 publications. We discuss the steps
21 required to develop this model system based on data resolution, sampling bias, and reporting quality. In addition,
22 we propose the following steps to guide efforts for a macroevolutionary model system for species interactions:
23 First, evaluating and expanding model system outcome measures to increase data resolution. Second, closing
24 knowledge gaps to address underreporting and sampling bias arising from limited human and financial resources.
25 Identifying phylogenetic and geographic targets, creating systematic overviews, enhancing scientific
26 collaborations, and avoiding data loss through awareness of predatory journal publications, can accelerate this
27 process. Third, standardising data reporting to increase reporting quality and to facilitate data accessibility.

28

29 **Key words**

30 Cichlid parasites; Dactylogyridae; Monogenea; host-parasite network; taxonomic bias; data reporting.

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40 **Conflict of interest**

41 The authors declare that they have no conflict of interest.

42 **Author contributions**

43 AJCL had the idea for this review, and performed literature search and data analysis. AJCL and MPMV wrote this
44 article with input from TA, AP, and KS.

45 **A model system for parasite evolution: The blueprint**

46 Species interactions such as predators-prey or host-symbiont, and host-parasite interactions are a key factor of the
47 evolution of species communities (terHorst et al., 2018). Interactions are the foundation of ecosystems (Duffy et
48 al., 2007; terHorst et al., 2018) and a crucial aspect for conservation efforts in the light of global ecosystem
49 degradation (Heleno et al., 2012; Tylianakis et al., 2010). However, traditional evolutionary biologists have mainly
50 focused on evolutionary processes in single species or pairwise interactions instead of the complex interaction
51 networks that ecologists have investigated extensively (terHorst et al., 2018). Study systems specifically used to
52 investigate evolutionary processes in species-rich interaction networks remain scarce (Toju et al., 2017) despite
53 ‘eco-evolutionary feedbacks’, i.e. the interplay of ecological and evolutionary processes, recently becoming a
54 focus point of evolutionary research (Toju et al., 2017). The root of this issue lies in the grouping of the species
55 (A) by their ecosystem function in relation to the other set (B) (e.g. A as a ‘pollinator’, ‘predator’, ‘prey’, ‘host’,
56 ‘symbiont’, or ‘parasite’ of B). Bipartite networks, one of the most common representations of interacting species
57 communities, frequently involve two such sets of interacting species groups (e.g. set A and set B) (Harvey et al.,
58 2017) such as plants and pollinators (Campbell et al., 2011), plants and phytophagous insects (Nylin et al., 2018),
59 conifers and granivorous birds (Benkman et al., 2010), plants and mycorrhizal fungi (Selosse et al., 2006), or corals
60 and dinoflagellates (Ziegler et al., 2018). As a consequence, the evolutionary distances amongst the species is
61 considerably larger than if A and B each formed recently evolved monophyletic taxa (e.g. Wilson et al., 2012).
62 Established bipartite networks frequently lack the power to integrate eco-evolutionary feedbacks as the extensive
63 period of time passed since the last common ancestor means that a multitude of factors might have affected the
64 diversification process of a species set. These factors can generate high levels of uncertainty for the detection of
65 evolutionary patterns.

66 Recent species radiation or even co-radiations, i.e. two co-evolving species radiations (Losos & Mahler, 2010),
67 could present a solution (Benkman et al., 2010) as the number of factors driving diversification processes is likely
68 lower due to the relatively short time frame since the last common ancestor. Species radiations are characterised
69 by explosive diversification events resulting in species-rich assemblages, and in some cases, replicate radiations
70 (Rabosky & Glor, 2010; Salzburger et al., 2005). Some model radiations such as the Caribbean lizard genus *Anolis*
71 (Squamata: Dactyloidae) (Mahler et al., 2013), Darwin’s finches (Passeriformes: Thraupidae) (Grant, 1999), and
72 cichlid fishes (Teleostei: Cichlidae) (Seehausen, 2006), have been explored in depth during decades of expansive
73 ecological, evolutionary, developmental, and behavioural research (Abzhanov, 2010; Koblmüller et al., 2015;
74 Sanger et al., 2008). However, the interactions of these models have only been explored from an ecological
75 perspective including studies on feeding strategies (Hata et al., 2014; Heleno et al., 2013) and predatory pressure
76 (Pringle et al., 2019). Co-radiations in general have been the subject of few evolutionary study systems (fungal
77 endophytes: Saikkonen et al., 2016; phytophagous insects: Wilson et al., 2012). Species-poor interaction networks
78 remain the norm for ‘eco-evolutionary’ studies (Toju et al., 2017). A model system involving metazoan co-
79 radiations has not been established to date.

80 Obligate parasite radiations and their hosts could be one of the most promising candidates for a macroevolutionary
81 model system for species interactions as host-parasite interactions are frequently preserved in natural history
82 collections of the host (Harmon et al., 2019) due to the close nature of parasitic interactions. Such a model system
83 for macroevolutionary study could shed light on patterns of disease evolution (Harmon et al., 2019) such as

84 expansions and reductions of host ranges and geographical ranges (Hoberg & Brooks, 2008), e.g. Jorissen et al.
85 (2020) especially because metazoan parasites account for a significant portion of all animal species worldwide
86 (Poulin, 2014) and are key drivers of evolutionary and ecological processes in species-rich communities (Gómez
87 & Nichols, 2013; Lefèvre et al., 2009). Furthermore, parasites are the cause of an increasing number of diseases
88 in humans, livestock (Thompson, 2013), and crops (Ekroth et al., 2019), an effect that is amplified by the current
89 process of global environmental degradation (Brooks et al., 2019). But despite the pool of historical records and
90 the integral role of parasites in ecosystem health, biosecurity, and public health, only few bipartite host-parasite
91 networks have been investigated to date as the small body sizes of most parasites complicates identifying species
92 (de Meeûs et al., 2007). Similar to other bipartite network systems, proposed metazoan parasite models have
93 focused on microevolutionary questions to minimise model complexity, i.e. systems involving few species and
94 species interactions, such as the water flea–bacterium network with *Daphnia magna* Straus, 1820 and *Pasteuria*
95 *ramosa* Metchnikoff, 1888 (Bento et al., 2017), the mammal–cestode network with several mammalian hosts and
96 species of *Echinococcus* Rudolphi, 1801 (Thompson & Jenkins, 2014), the legume–pea aphid network with
97 leguminous plants and several pea aphid subspecies (Peccoud & Simon, 2010), and several fish–flatworm networks
98 involving gyrodactylid monogenean parasites (Bakke et al., 2007), e.g. the salmonid–*Gyrodactylus* (Ziętara &
99 Lumme, 2002) or goby–*Gyrodactylus* (Huyse & Volckaert, 2005) networks, but see species-rich plant parasite
100 networks, e.g. Wilson et al. (2012). To investigate the macroevolution of host-parasite interactions, the flatworm
101 parasites belonging to *Cichlidogyrus* Paperna, 1960 (Platyhelminthes: Monogenea, Dactylogyridae) [including
102 *Scutogyrus* Pariselle & Euzet 1995 (Platyhelminthes: Monogenea, Dactylogyridae) (Wu et al., 2007)] infecting the
103 gills of mainly African cichlid fish have been proposed as a study system for metazoan parasites (Pariselle et al.,
104 2003; Vanhove et al., 2016). Arguments expressed in favour of a cichlid–*Cichlidogyrus* model system can be
105 summarised as follows [according to the model system criteria of Kellog & Shaffer (1993)]:

- 106 • *Variability*. Monogeneans are the most species-rich group of cichlid parasites alongside digeneans (Vanhove
107 et al., 2016) with *Cichlidogyrus* being the most species-rich African freshwater fish parasite genus (Kuchta et
108 al., 2018). The cichlid–*Cichlidogyrus* network currently comprises 138 parasite species infecting fish species
109 belonging to fifteen cichlid tribes and two non-cichlid families resulting in 416 host-parasite interactions
110 (Online Resource 1). This known species diversity is higher than the diversity of all parasitic nematodes in
111 African fish (Moravec, 2019) and provides the cichlid–*Cichlidogyrus* system with an advantage over other
112 systems involving host model systems with a less diverse parasite fauna such as helminth parasites of anoles
113 (Falk & Perkins, 2015) or feather mites of Darwin’s finches (Bulgarella & Palma, 2017).
- 114 • *Tractability*. Sample availability and, therefore, data availability and quality of host parameters are high due
115 to the model system status of African cichlids (Kornfield & Smith, 2000) rooted in the extensive library of
116 scientific publications (see Koblmüller et al. 2015). In addition, a simple life cycle of the parasites in the
117 absence of intermediate hosts, a lower infection and attachment site diversity of fish gills as opposed to internal
118 organs (Bellay et al., 2015), and a predicted neutral effect on the natural selection of the host species (Pariselle
119 et al., 2003), minimise the number of parasite-related parameters (Pariselle et al., 2003). The gill flap protects
120 the gill parasites from physical damage during the host sampling process that can potentially skew the recorded
121 infection levels of a host specimen (Kvach et al., 2016). Furthermore, the ectoparasitic lifestyle protects
122 samples from degradation as the parasite is instantly exposed to any fixative media (e.g. formaldehyde or
123 ethanol), in which the host specimens are fixed, stored, and preserved. In contrast, the preservation of

124 endoparasites is delayed by the host's body, which the fixative has to penetrate to reach the infection site. This
125 delay can result in a deteriorated state of morphological characteristics and DNA molecules (Tkach et al.,
126 2019) unless the parasite samples are extracted immediately after the death of the host.

- 127 • *Accessibility.* Due to the sympatric speciation of cichlids in Africa, most host species are found in the
128 biodiversity hotspots of the Great East African Lakes. This focused geographical range (Pariselle et al., 2003)
129 and the availability of host specimens from natural-history collections worldwide (Harmon et al., 2019)
130 increase the accessibility to host samples.
- 131 • *Representativeness.* As flatworms (Platyhelminthes), species of *Cichlidogyrus* are representatives of one of
132 the groups with the highest parasite diversity alongside arthropods (Poulin & Morand, 2004). In addition, their
133 cichlid hosts have the biodiversity hotspot in the East African Lakes. This hotspot is also a global biodiversity
134 hotspot. Hence, understanding interactions of cichlids and non-cichlids in multispecies networks should be
135 considered key for conserving diverse species communities as a whole especially because climate change
136 (Cohen et al., 2016; Olaka et al., 2010), biological invasion (Witte et al., 2013), pollution (Odada et al., 2003),
137 and overfishing (Kayanda et al., 2010) threaten the unique ecosystems of the East African lakes. Only few
138 studies have investigated cichlid–non cichlid interactions such as with flatworm parasites (Mendlová &
139 Šimková, 2014), algae (Hata et al., 2014) and hydrozoans (Bose et al., 2019). While no replicate radiations
140 have been described to date, natural (Messu Mandeng et al., 2015) and invasion-induced (Jorissen et al.,
141 2018b; Roche et al., 2010; Šimková et al., 2019) host shifts have occurred in multiple cases. Investigating
142 these events could provide insight into pattern of disease evolution as novel pathogens frequently emerge from
143 host shifts (Woolhouse et al., 2005).

144 Despite these arguments, the model system quality of the cichlid-*Cichlidogyrus* network has never been evaluated
145 systematically. Numerous studies have produced a broad but scattered insight into the diversity, ecology,
146 epidemiology, and evolution of cichlid-monogenean interactions (see Vanhove et al., 2016). Data of
147 morphological, phylogenetic, and ecological studies are published with varying quality. Therefore, this review will
148 assess shortcomings of literature data and discuss and propose solutions of three common issues regarding these
149 data. We will first give an overview on the issues and solutions related to data resolution, i.e. which measurements
150 and markers are being used and could be used in the future as ‘the model system toolbox’. Second, we will
151 characterise the sampling bias and data loss, identify causes, and propose a range of solutions for publications on
152 cichlid parasites to ‘create a stable model system architecture’. Third, we will assess reporting quality and propose
153 a framework to optimise reporting quality of data, the ‘foundations of model system architecture’. We will estimate
154 whether and how improvements in these areas can enhance model system aspects of the two–trophic level network.
155 This review aims to guide future efforts to develop a bipartite model system to analyse macroevolutionary
156 mechanisms of host-parasite interaction networks.

157 **Measurements and markers: The model system toolbox**

158 A key tool for measuring and describing mechanisms in a model system are endpoints. Endpoints are outcome
159 measures of a model system commonly applied in ecosystem health assessment (Cairns et al., 1992). If we apply
160 this concept to the cichlid-*Cichlidogyrus* system, morphological, ecological, and genetic parameters included in
161 typical reports of *Cichlidogyrus* spp. could constitute model endpoints. However, the parameters commonly
162 considered in these reports fail to provide enough data resolution as they are limited to a few morphological

163 characteristics (see Fig. 1 for a typical drawing), genetic markers, and ecological measures. To build the cichlid-
164 *Cichlidogyrus* model system and increase the system's tractability, the systematic and ecological informativeness
165 of established endpoints should be assessed and novel endpoints should be established as follows.

166 First, past studies have characterised species of *Cichlidogyrus* and other monogenean species mainly with
167 morphometric measures of the sclerotised structures in the attachment (haptor) and reproductive organs (see Fig.
168 1 for measures included in typical species descriptions and standard terminology), e.g. in Fannes et al. (2017), and
169 few partial nuclear and mitochondrial DNA markers, e.g. in Igeh et al. (2017). Some haptor measures coincide
170 with phylogenetic clades (Vignon et al., 2011) as demonstrated by some studies using ancestral states
171 reconstruction of morphological characters of species of *Cichlidogyrus* (Mendlová et al., 2012; Vignon et al.,
172 2011). However, these results appear questionable as undersampling of specimens and distortion of those
173 morphological characters as a result of sample fixation (Fankoua et al., 2017; Vignon, 2011) might produce biased
174 results and systematic errors. Some measures might also be highly correlated, i.e. multicollinear, and, therefore,
175 redundant (Vignon, 2011). To maximise systematic and ecological informativeness, the scale of undersampling,
176 distortion, and redundancy of measures should be quantified through a statistical assessment of these
177 morphological characteristics. This assessment could increase statistical power of data analyses in the future.
178 Additional measures should be added to include shape variations that are not covered by current methods, e.g. the
179 width and shaft lengths of the hooks (Geraerts et al., 2020), the membranous extension of the ventral bar (Mendlová
180 et al., 2012), and the interspecific shape variations of penis and accessory piece in the male copulatory organ
181 (Geraerts et al., 2020). Similar to a previous study on the anchor shape variations in species of the genus
182 *Ligophorus* Euzet & Suriano, 1977 (Rodríguez-González et al., 2017), geometric morphometric analyses could be
183 used to maximise the phylogenetic signal of the measures. Furthermore, new imaging and staining techniques such
184 as three-dimensional images resulting from scanning electron microscopy (Fannes et al., 2015), confocal
185 microscopy (Galli et al., 2006; Marchiori et al., 2015; Rossin et al., 2017) or soft tissue staining (Mair et al., 1998)
186 might enhance morphological and anatomical descriptions and, thus, increase data resolution.

187 Second, most phylogenetic analyses have been based on few genetic markers including the partial 28S and 18S
188 nuclear ribosomal DNA, partial internal transcribed spacer 1 (ITS1) region of nuclear ribosomal DNA, and
189 mitochondrial cytochrome *c* oxidase subunit I (COI) DNA regions. These studies have resulted in mostly modest
190 support values (Messu Mandeng et al., 2015; Mendlová et al., 2012) leaving questions on the systematics of
191 *Cichlidogyrus* unanswered. Furthermore, the systematic application of other genetic markers such as 16S nuclear
192 ribosomal DNA (Plaisance et al., 2005), and the histone 3 (H3), the elongation factor 1 α (EF1 α) (Perkins et al.,
193 2009), and the mitochondrial cytochrome *c* oxidase subunit II (COII) DNA region (Bueno-Silva & Boeger, 2014)
194 remains restricted to certain other monogenean taxa. New mitochondrial and nuclear DNA markers might soon
195 increase resolution of phylogenetic and population-level analyses in future studies especially because of recent
196 advances in assembling DNA sequences of specimens of *Cichlidogyrus* resulting from next-generation sequencing
197 techniques (Vanhove et al. 2018).

198 Third, infection parameters, e.g. prevalence, i.e. ratio of infected hosts, and infection intensity, i.e. parasites per
199 infected hosts, are the most common quantifiers of cichlid-*Cichlidogyrus* interactions as these parameters are
200 essential and accessible quantifiers of parasite ecology (Poulin, 2006). However, beyond infection levels,
201 ecological parameters are mostly absent from typical taxonomic reports despite several studies demonstrating that

202 host size and sex (Akoll et al., 2012), seasonality (Akoll et al., 2012; Hirazawa et al., 2010), and inorganic and
203 organic pollutants (Poléo et al., 2004; Sanchez- Ramirez et al., 2007; Šebelová et al., 2002) affect the physiology
204 and host interactions of flatworm parasites. To improve current data sets, related to the host (e.g. Akoll et al., 2012)
205 or the immediate environment (e.g. Fannes et al., 2017; Huyse et al., 2004) should be included in full in the
206 respective publication if the data collection does not inflate the research effort as even few observations can provide
207 information for meta-analyses. Additional host endpoints will further increase the data resolution of the proposed
208 model system, and can be found in studies focussing on the pathology (e.g. Igeh & Avenant- Oldewage, 2020)
209 and immunology (e.g. Zhi et al., 2018) of monogenean infections or the effects of human activity on the disease
210 ecology (e.g. Madanire-Moyo et al., 2012). More insight into the health of cichlids is needed as infrastructure and
211 human resources addressing the biosecurity of these species remains poor (Opiyo et al., 2018) with little disease
212 monitoring of cultured (Opiyo et al., 2018) and wild (Mugimba et al., 2018) populations of ‘tilapias’ despite the
213 economic relevance of these cichlids for fisheries and aquaculture in Africa, Asia, and the Americas (FAO, 2019).
214 We also recommend the deposition of ‘symbiotypes’, i.e. host vouchers of parasite types, in natural history
215 collection because similar to other diseases (Bradley et al., 2020) this deposition would have the advantage that
216 many host parameters are preserved in these collections long-term. For environmental parameters, *in-vitro*
217 culturing methods, i.e. cultures without living hosts, might facilitate investigating the effects of environmental
218 factors in controlled laboratory environments in the near future as these methods are becoming available for an
219 increasing number of parasite species (Hutson et al., 2018), some of which are relatively closely related to
220 monogeneans, e.g. the cestode *Schistocephalus solidus* Müller, 1776 (Wedekind et al., 1998), or also have a similar
221 ectoparasitic lifestyle, e.g. gnathiid isopods (Grutter et al., 2020).

222 **Sampling bias: Creating a stable model system architecture**

223 Cichlid-*Cichlidogyrus* interactions constitute the most extensively described host-parasite network based on an
224 adaptive radiation of host species. Decades of research (Online Resource 2) have provided a valuable insight into
225 the biodiversity and ecology of these parasites. Yet the knowledge on the cichlid parasite diversity remains
226 incomplete, biased, and insufficient for conclusive genus-wide evolutionary analyses as the discovery of new
227 species is constrained by limited human and financial resources. The dominance of a few researchers in novel
228 reports of parasite-host interactions (Fig. 2) and a rapid increase of new species of *Cichlidogyrus* highlight these
229 constraints. Since the last taxonomic revision of dactylogyridean parasites of African cichlids (Pariselle & Euzet,
230 2009), 59 new species of *Cichlidogyrus* have been recorded, a 75% increase. Some of these species have been
231 reported from cichlid hosts belonging to seven previously neglected tribes (Online Resource 2). At the same time,
232 new discoveries of host-parasite interactions have continued to increase exponentially, i.e. the discovery rate shows
233 no sign of deceleration and appears unrestrained due to the extensive numbers of interactions remaining to be
234 found. However, this increase (generalised linear model with negative binomial distribution, $\chi^2 = 243.41$, $df = 1$,
235 $p < 0.001$) (see Fig. 2) suggests that the discovery of cichlid parasite species and host-parasite interactions remains
236 at the characteristic early exponential stage of a typical cumulative species discovery curve (Bebber et al., 2007).
237 Therefore, estimating the total number of extant interactions is currently impossible. Many more interactions could
238 be discovered in the future if restricting factors are addressed.

239 One of these restricting factors is research capacity. Like other biodiverse regions (Fisher & Christopher, 2007),
240 challenging political and economic conditions surrounding the East and Central African cichlid biodiversity

241 hotspots continue to limit funding opportunities and the access to sampling locations despite a growing scientific
242 impact of publications from developing countries (see Gonzalez-Brambila et al., 2016). For researchers in Africa,
243 these restrictions inhibit the access to specimens in biorepositories, e.g. natural history collections, as samples are
244 frequently moved to and stored in developed countries (The H3Africa Consortium, 2014), see ‘parachute science’
245 (Barber et al., 2014). Financial limitations also restrict the access to scientific literature as commercial interests of
246 publishers (Fuchs & Sandoval, 2013) result in high access fees for readers or, alternatively, high publication fees
247 for the authors of open access articles (Veríssimo et al., 2020). These disparities between high and low-income
248 countries cause structural and financial dependencies (Dahdouh-Guebas et al., 2003) and inhibit the growth of
249 research capacity as demonstrated by the higher percentage of foreign institutions in the collaborative networks of
250 research institutions in developing countries compared to industrialised nations (Banerjee, 2017). Consequently,
251 societal preferences in both developing and developed countries drive a positive and taxonomic publication bias:
252 positive and novel results are favoured over negative and repeat observations whereas cichlid hosts relevant to
253 consumption, i.e. species belonging to the tilapia-like tribes Coptodonini Dunz and Schlieven, 2013 (22% of
254 publications) and Oreochromini Dunz and Schlieven, 2010 (65% of publications), are favoured over less
255 economically relevant hosts belonging to other tribes (33% of publications) (Fig. 3; Online Resource 2). This
256 research impact-driven approach, which focuses on novelty and economic relevance, is reminiscent of publication
257 (Song et al., 2010) and taxonomic biases (Troutet et al., 2017) in global biodiversity data and research biases in
258 other fields such as conservation research (Trimble & Van Aarde, 2010; Ford et al., 2017), epidemiology (Siddiqi,
259 2011), and invasion ecology (Pyšek et al., 2008). Because of this ‘bias’, most evolutionary research concerning
260 *Cichlidogyrus* remains purely descriptive or limited to a subset of species of *Cichlidogyrus*, e.g. exclusively to
261 parasites infecting tropheine (Vanhove et al., 2015) or bathybatine (Kmentová et al., 2016) cichlids. Taxonomic
262 biases in genus-wide studies can lead to misleading results as the ecology of species of *Cichlidogyrus* can vary in
263 the same ecosystem depending on the host species (Kmentová et al., 2016). Thus, an observed correlation could
264 reflect a sampling bias rather than causation, e.g. Mendlová & Šimková (2014) detected a correlation of parental
265 care behaviour and host specificity because of a study bias towards cichlid species expressing a specific type of
266 parental care, namely mouth brooding.

267 Both publication and taxonomic biases are amplified by data that continue to be lost through low quality
268 publishing. In the light of the exponential growth of scientific publishing (Bornmann & Mutz, 2015), assuring
269 research quality has become a discussion that exceeds the scope of this review. However, one prominent
270 phenomenon arising from publication pressure in taxonomic research (e.g. Raghavan et al., 2014) is predatory
271 publishing, i.e. exploitative publications that demand high fees with no attached peer-review process (Beall, 2013).
272 Data arising from predatory publishing are effectively ‘lost science’ (Clark & Smith, 2015) as research quality
273 cannot be guaranteed. Cichlid research forms no exemption as predatory journals specifically target researchers
274 from low-income countries (Xia et al., 2015), where most cichlid species are sampled. Predatory publishing can
275 further boost the stigma for publications from these countries, which can already face a lack of valorisation (Lee
276 et al., 2013). A potential bias of peer-reviews against early-career researchers (Lee et al., 2013) from low-income
277 countries [see ‘matata effect’ in Van der Stocken et al. (2016)] might enhance the success of this targeting strategy.

278 To address sampling bias and data loss, closing knowledge gaps is key. Recent studies on the parasites of host
279 species belonging to the underrepresented cichlid tribes including Bathybatini Poll, 1986 (Kmentová et al., 2016);
280 Cyprichromini Poll, 1986; Ectodini Poll, 1986; Eretmodini Poll, 1986 (Rahmouni et al., 2017); Haplochromini

281 Poll, 1986 (Van Steenberg et al., 2015); Hemichromini Hoedeman, 1947; and Tylochromini Hoedeman, 1947
282 (Muterezi Bukinga et al., 2012; Jorissen et al., 2018a; Pariselle et al., 2014; Rahmouni et al., 2018) have expanded
283 the taxonomic coverage concerning examined host specimens of *Cichlidogyrus*. However, knowledge gaps can be
284 addressed more systematically taking the following approaches:

- 285 • *Targeting*: Databases of literature data (Online Resource 2) with taxonomic (Online Resource 1) and
286 geographical inventories of reported samples should be created, and the deposition of voucher specimens in
287 biorepositories close to the sample locations should be encouraged to increase data access and availability in
288 low-income countries. These measures could outline gaps in host taxonomic and geographic coverage (Jones
289 et al., 2009) and support decision-making on taxonomic and geographic targets (Stephenson et al., 2017)
290 similar to the handbooks for the general parasite diversity (see Scholz et al., 2018) and nematode parasite
291 diversity (Moravec, 2019) of African fish.
- 292 • *Systematic revision*: The 2009 systematic revision of *Cichlidogyrus* and the identification key (Pariselle &
293 Euzet, 2009) should be updated to facilitate identifying species despite the relatively small sizes of these
294 parasites.
- 295 • *Collaboration*: ‘Parachute’ science, i.e. research in developing countries that side-lines local scientific
296 communities (Barber et al., 2014), can be avoided through North-South and South-South collaborative
297 networks involving African institutions (Chu et al., 2014; Klopper et al., 2002). North-South collaboration
298 can, e.g., facilitate the access for local academics to distant learning resources to make acquiring research
299 skills more affordable (Chu et al., 2014). Promoting South-South collaborations through an improved research
300 infrastructure for genomic research including African-based biorepositories and standardised protocols for
301 sharing these resources might also reverse the ‘brain drain’ of African experts to high-income countries (The
302 H3Africa Consortium, 2014). These networks could enable the whole research community to tap into local
303 scientific and traditional knowledge (Drew, 2005) while increasing the academic capacity in the region and
304 the whole field (Barber et al., 2014; Van der Stocken et al., 2016). The need for capacity development in
305 monitoring and identifying parasites in the Global South has also been observed in a One Health or EcoHealth
306 context (Keune et al., 2017).
- 307 • *Awareness building*: Increasing awareness of predatory publishing (Beall, 2013) and reforming the current
308 systems for evaluating and funding science (Edwards & Roy, 2017) are ideas to reduce data loss. Systemic
309 changes could include enforcing ethical codes in peer-review reports to avoid ‘bad faith’ reviews (Atjonen,
310 2018), double-blind (Tomkins et al., 2017) or open review (Ross-Hellauer, 2017) processes to address
311 reviewer biases (see references for more detailed information), and publications in academic journals that
312 follow Fair Open-Access principles (Verissimo et al., 2020) to ease financial restrictions in low-income
313 countries.

314 In conclusion, taking a more targeted, systematic, and awareness-building approach will enhance the tractability
315 of the *Cichlidogyrus*-cichlid network whereas the collaborative aspect might additionally enhance model
316 accessibility due to an increased academic capacity for parasitological research in Africa.

317 **A standard for reporting data: The foundations of model system architecture**

318 As data from cichlid gill parasites reach back 80 years, data reporting has undergone considerable changes from
319 basic taxonomic descriptions in the 1930s (Markewitsch, 1934) to modern data sheets and DNA sequence data
320 deposited in online databases (e.g. Igeh et al., 2017). However, data accessibility varies across current publications
321 due to barriers working against data dissemination such as varying reporting quality. Out of 163 publications on
322 *Cichlidogyrus* and *Scutogyrus*, approximately a third report prevalence data and somewhat fewer report infection
323 intensity data across the three most studied tribes (Fig. 3, Online Resource 2). This lack of data accessibility should
324 be addressed by increasing reporting quality in future publications and, thus, tractability of the network. We
325 propose a standardised reporting framework (Table 1).

326 Varying data accessibility inhibits data dissemination (Molloy, 2011) and prolongs data extraction (Dasu &
327 Johnson, 2003). For cichlid parasite literature, data extraction frequently requires manual copying of hidden data
328 from non-modifiable formats such as in running text or portable document format (pdf) tables. Furthermore,
329 varying reporting quality hampers data analyses due to missing data and laborious calculations of parameters from
330 raw data. In taxonomic publications on cichlid parasites, data ranges (Dossou, 1982; Paperna & Thurston, 1969;
331 Rahmouni et al., 2018) and qualitative data (Pariselle et al., 2015) for ecological and morphological parameters
332 are traditionally accepted for describing and, consequently, identifying species. Hence, infection parameters, and
333 mean values of infection parameters and morphometric data are frequently absent despite these quantifiers being
334 essential and accessible descriptors of parasite ecology (Poulin, 2006). Epidemiological publications include
335 infection parameters prominently (Firmat et al., 2016; Otachi et al., 2014) but frequently lack species names as the
336 species are only identified on genus level due to the focus of these studies on the disease ecology rather than the
337 parasite biodiversity.

338 A standardised reporting framework can improve data reporting quality (see Table 1). Reported morphological
339 and ecological measurements ought to include mean values and descriptors of data distribution such as minimum,
340 maximum, standard deviation, and sample size, and descriptors of host-parasite interactions, i.e. infection
341 parameters, including examined host individuals, infected host individuals, and mean parasite load. Data ought to
342 be reported in ‘tidy’ data tables (Wickham, 2014), i.e. one variable per column and one observation per row, saved
343 in open formats such as comma-separated values (csv) or tab-delimited text files (txt). These raw data should be
344 openly accessible in supplementary files or, preferably, public repositories (Molloy, 2011) to support the data
345 summarised in the main article. We recommend this framework for data reporting to increase tractability of the
346 network by optimising data accessibility for future studies.

347 **Work in progress: A conclusion**

348 We demonstrated that cichlid fishes and the gill-infecting flatworm parasites belonging to *Cichlidogyrus* are
349 uniquely suitable for being bipartite macroevolutionary model system due to the presence of historical collections
350 and the range of different fields that have started exploring these interactions from different perspectives including
351 evolutionary, epidemiological, ecological, and genomic research. The combination a species-rich clade of parasites
352 with a single-host lifecycle and a host diversity that has been studied for decades are unique for a metazoan parasite
353 model. However, the low data resolution, the sampling bias towards economically relevant host species, and the
354 inconsistent data quality demonstrate that many of these fields have only recently started to explore the parasites
355 of cichlids. Therefore, many model system criteria cannot currently be met but the present model aspects might be
356 enhanced in the future if the measures are taken to address shortcomings with additional research. To do so, we

357 proposed a range of approaches, some of which have taken first steps in recent years including genomics and ‘tidy’
358 data, and others which appear futuristic such as the possibility of *in vitro* cultures. With these proposals, we have
359 aimed to highlight the different perspectives on host-parasite networks or species interactions in general ranging
360 from taxonomic, evolutionary, ecological standpoints to pathological, epidemiological, immunological, and data
361 science perspectives. We also addressed underlying social aspects of parasitological research in low-income
362 countries. These systemic problems of scientific publishing and research in the Global South might be discussed
363 more thoroughly elsewhere. However, we have demonstrated that these issues often lie at the root of many
364 ‘scientific’ problems at a second glance.

365 Overall, this review has highlighted the importance of data science to building a macroevolutionary model system
366 for species interactions. All bipartite systems ranging from symbiotic and predator-prey interactions to viral and
367 bacterial diseases are at their core networks of interacting species. Nonetheless, ‘eco-evolutionary feedback’
368 studies with bipartite models have lagged behind their ‘unipartite’ counterparts. With this review, we would like
369 to encourage researchers working on other species interactions to evaluate data resolution, sampling biases, and
370 reporting quality of the respective system to detect shortcomings. Naturally, priorities on the data that are collected
371 and published differ amongst taxonomists, evolutionary biologists, epidemiologists, and ecologists. For example,
372 taxonomic and epidemiological perspectives have shaped the reporting of metazoan parasites and their animal
373 hosts in flatworm research, i.e. the data are presented to meet the requirements of the field. Yet any raw data should
374 be published even if those are not directly relevant to the respective publication as this simple effort can create
375 opportunities for researchers in high and low-income countries. Researchers should also consider adding their raw
376 data to openly accessible databases. To characterise a host-parasite system, morphological data, interaction data,
377 and DNA sequences data can be uploaded to MorphoBank, Global Biotic Interactions (GloBI), and NCBI
378 GenBank. An open exchange of data could multiply the impact of study systems and increase the interconnectivity
379 of research groups and fields.

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764 **List of figures**

765 **Figure 1** Schematic overview of typical morphological structures and measurements in reports of *Cichlidogyrus*
766 spp. A Overview of organism with haptor sclerites and male copulatory organ. B Anchor. C Hook. D. Ventral bar.
767 E Dorsal bar. F Male copulatory organ. Measurements: a, anchor total length; b, anchor blade length; c, anchor
768 shaft length; d, anchor guard length; e, anchor point length; U1-U7, hook lengths; w, bar width; x, bar length; h,
769 auricle length; y, distance between auricles; AP, accessory piece length; Pe, penis length; He, heel length;
770 sclerotised vagina not included as only observed in some species

771 **Figure 2** Cumulative count of host-parasite interactions reported between 1934 and 2019 for species of
772 *Cichlidogyrus* and *Scutogyrus*, grouped by first authors of publications first recording these combinations. The
773 increase of discovered combinations mainly relates to the works of I. Paperna and A. Pariselle. Furthermore, the
774 exponential growth of the discovery curve (grey) indicates that these discoveries remain at an early stage,
775 independent of overall species number (Bebber et al., 2007). Discoveries are not independent as new accounts are
776 published in groups (Bebber et al., 2007). To account for the overdispersion caused by this dependency, a discovery
777 curve was fitted using a generalised linear model with a negative binomial distribution ($\chi^2 = 243.41$, $df = 1$, $p <$
778 0.001) using the ‘MASS’ package (Venables & Ripley, 2002) in the statistical software R v3.6.1 (R Core Team,
779 2019). Graphing was performed using the ‘ggplot2’ package from R (Wickham, 2016)

780 **Figure 3** Number of peer-reviewed publication (until July 2020) on *Cichlidogyrus* and *Scutogyrus* infecting the
781 three most examined cichlid tribes (All), and number of those publications reporting infection parameters such as
782 prevalence (P) and infection intensity (I). *Pooled data from other cichlid tribes also include *Cichlidogyrus* spp.
783 infecting non-cichlids. Graphing was performed using the ‘ggplot2’ package (Wickham, 2016) in the statistical
784 software R v3.6.1 (R Core Team, 2019)

785

786 **List of electronic supplementary material**

787 **Online Resource 1** List of first records of host-parasite interactions of parasitic flatworms belonging to
788 *Cichlidogyrus* (Dactylogyridae, Monogenea) and their cichlid (Cichlidae, Teleostei) and non-cichlid host fishes.

789 **Online Resource 2** List of records of host-parasite interactions of parasitic flatworms belonging to *Cichlidogyrus*
790 (Dactylogyridae, Monogenea) and their cichlid (Cichlidae, Teleostei) and non-cichlid host fishes in peer-reviewed
791 publications, and presence of infection parameters for those records.

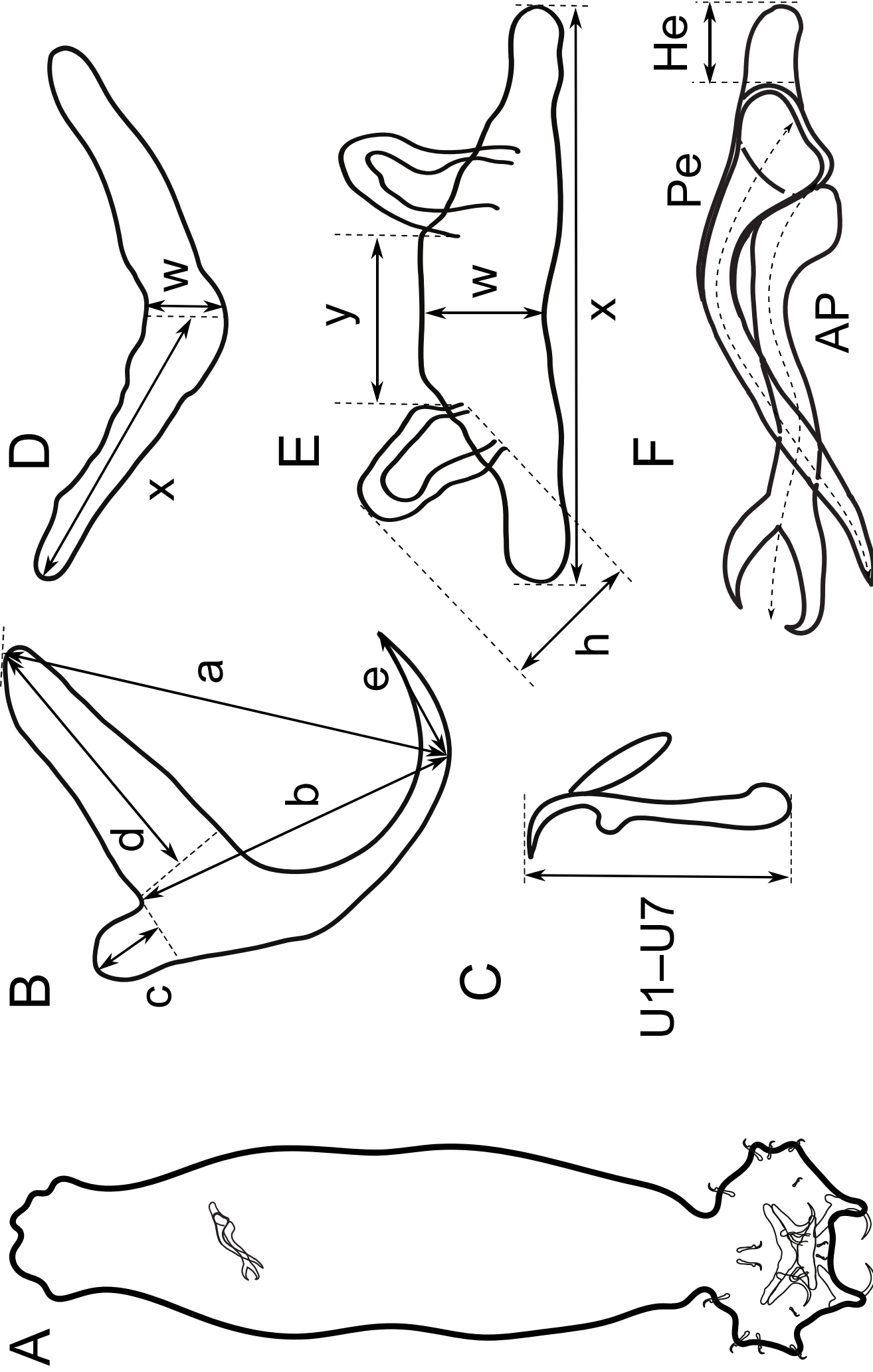


Figure 2

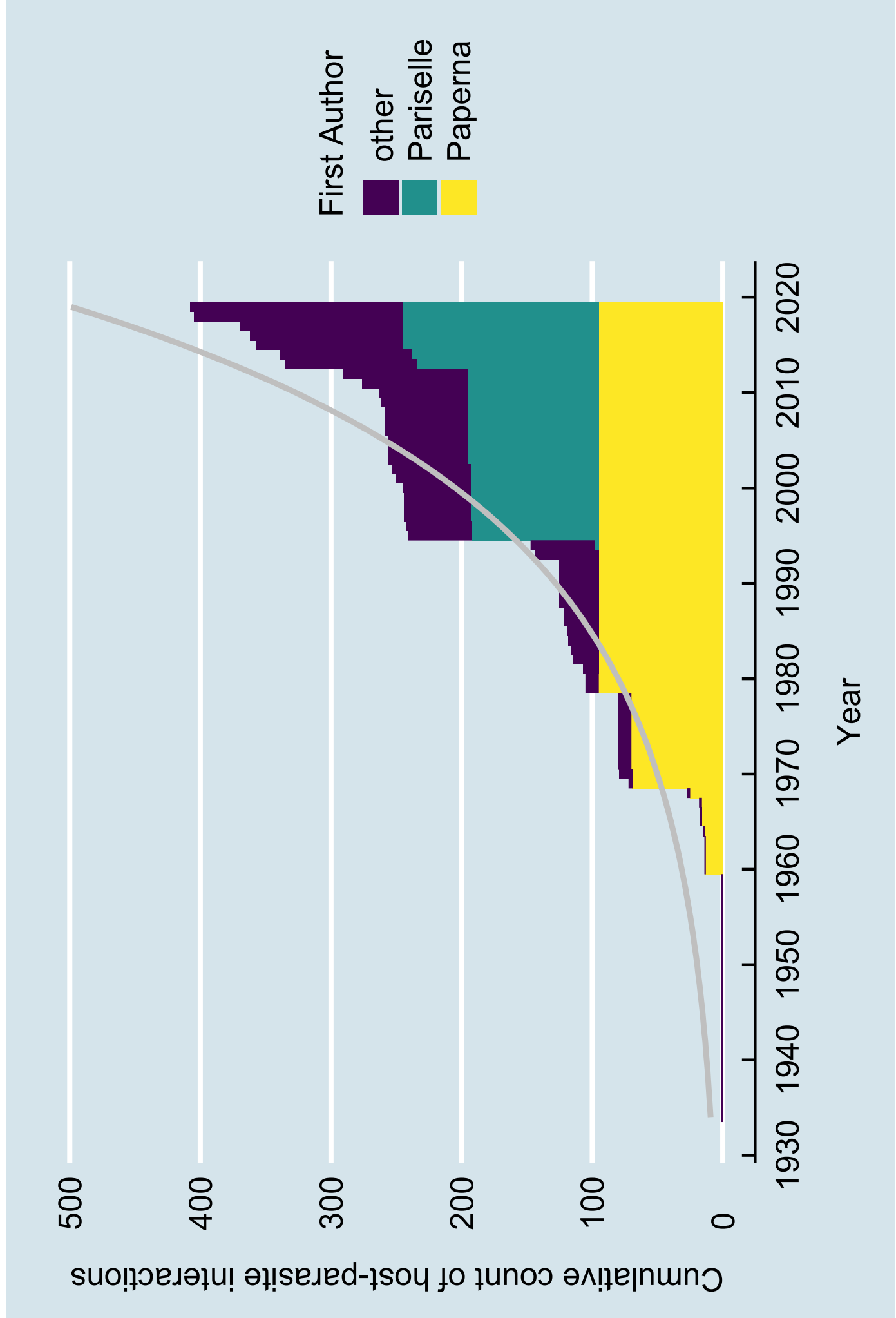


Figure 3

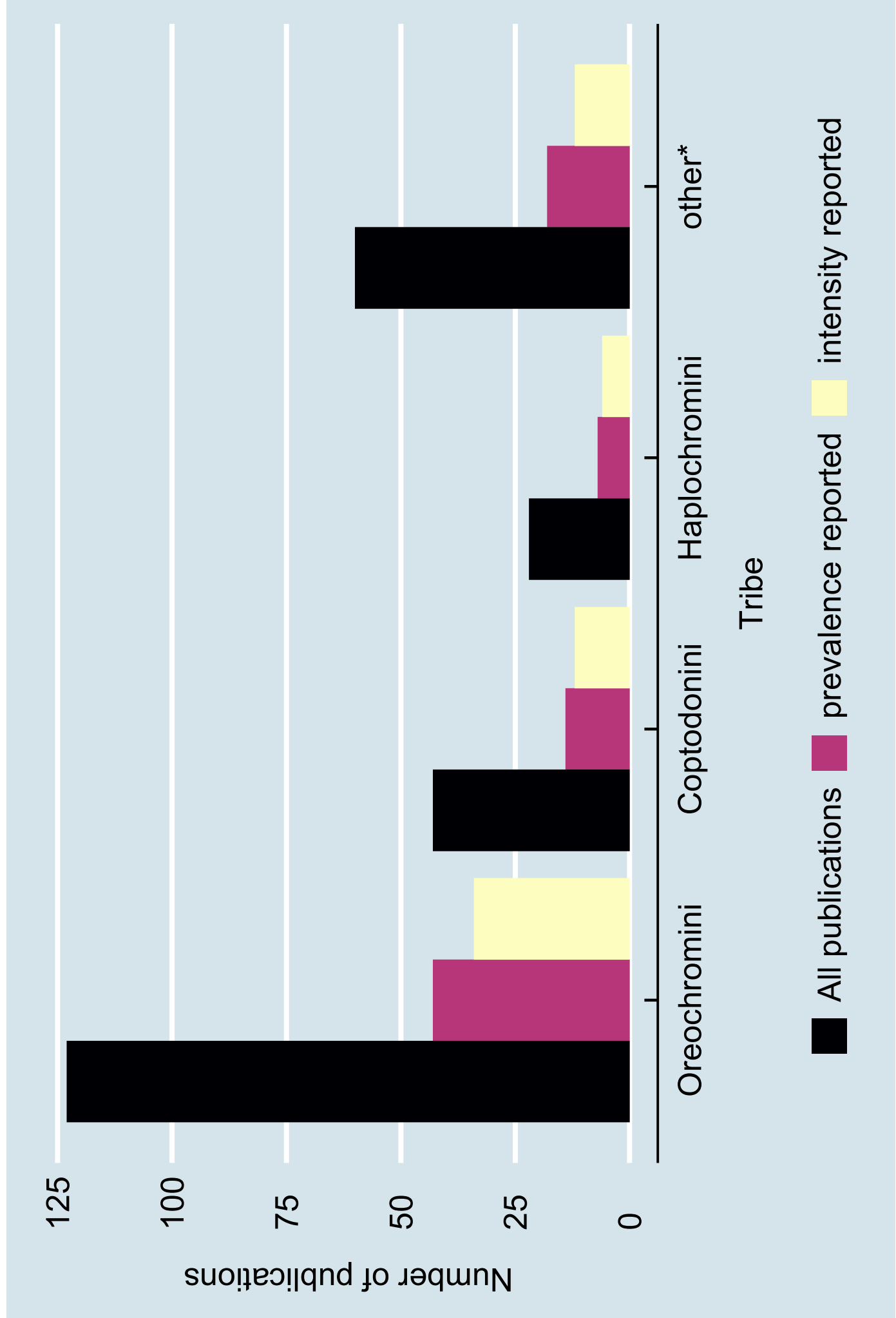


Table 1 Proposed standardised framework for reporting host-parasite network data, common reporting practice (examples in italics), and consequences of the common practice for data analysis: enhanced number and data formatting can increase reporting quality in future publications.

Standardised framework	Common practice	Consequence of common practice
Number formatting		Lost data: Usefulness for quantitative analyses limited
<ul style="list-style-type: none"> All measurements incl. mean and data distribution ($n > 30$) <i>mean: 23 μm</i> <i>min: 20 μm</i> <i>max: 25 μm</i> <i>standard deviation: 2.4 μm</i> Fully reported infection parameters <i>examined hosts: 29</i> <i>infected hosts: 10</i> <i>parasites (total): 45</i> Host parameters (with mean and data distribution if applicable) <i>Size: 15 cm</i> <i>Sex: Male/female</i> Physico-chemical parameters (with mean and data distribution if applicable) <i>e.g. temperature: 29</i> 	<ul style="list-style-type: none"> Incomplete reporting, e.g. as range <i>17-23 μm</i> Qualitative reporting <i>"We found parasite A infecting host B."</i> Incomplete reporting <i>Intensity: 1-20</i> No reporting No reporting 	
Data formatting		Time intensive extraction of data
<ul style="list-style-type: none"> Data in tables Tidy data table (1 variable per column, 1 observation per row) <i>mean: 23 μm</i> <i>min: 20 μm</i> <i>max: 25 μm</i> <i>standard deviation: 2.4 μm</i> Open formats in supplementary file or public repositories <i>'table.csv'</i> <i>'table.txt'</i> 	<ul style="list-style-type: none"> Data in running text 'Messy' data table <i>(20-)21-24(-25) μm</i> Non-modifiable formats <i>'article.pdf'</i> 	