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4	Title: Daily temperature variation magnifies the toxicity of a mixture consisting of a chemical
5	pesticide and a biopesticide in a vector mosquito
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Abstract: While many studies on the toxicity of pesticides looked at the effects of a higher mean 27 temperature, effects of the realistic scenario of daily temperature variation are understudied. 28 Moreover, despite the increasing interest for the toxicity of pesticide mixtures how this is 29 influenced by temperature has been largely ignored. We tested whether daily temperature 30 variation (DTV) magnifies the toxicity of two pesticides with a different mode of action, the 31 32 organophosphate pesticide chlorpyrifos (CPF) and the biopesticide Bacillus thuringiensis var. *israelensis* (Bti), and of their mixture in the vector mosquito *Culex pipiens*. Single exposure to 33 34 CPF and Bti increased mortality and reduced female development time, and exposure to CPF 35 also increased female wing length. DTV was not lethal and did not change the toxicity of the individual pesticides. Yet, a key novel finding was that high DTV increased the mortality of the 36 mixture by changing the interaction between both pesticides from additive to synergistic. Given 37 that in nature daily temperature variation is omnipresent, this is important both for vector control 38 39 and for ecological risk assessment. The higher toxicity of the mixture at high DTV compared to 40 the typically used constant test temperatures in the laboratory urges caution when evaluating the environmental impact of pesticide mixtures. 41

42 Capsule: We provide the first evidence that daily temperature variation can magnify the toxicity43 of a pesticide mixture by changing the interaction type between the pesticides.

44 Key words: antagonism; climate change; global warming; interaction effect; synergism;

45 integrated vector management

46 Highlights:

• The influence of temperature on the toxicity of mixtures has largely been ignored

• Mosquitoes were exposed to two pesticides under daily temperature variation (DTV)

• DTV did not increase the toxicity of the single pesticides

- DTV increased the toxicity of the pesticide mixture
- DTV is omnipresent in nature, hence important for risk assessment of mixtures

Abbreviations: <sup>1</sup>Bti - *Bacillus thuringiensis* var. *israelensis*, <sup>2</sup>CPF – chlorpyrifos, <sup>3</sup>DTV – daily
temperature variation, <sup>4</sup>DTV-0 = constant temperature, <sup>5</sup>DTV-7 = daily temperature variation of
7 °C, <sup>6</sup>DTV-14 = daily temperature variation of 14 °C, <sup>6</sup>IVM – integrated vector management,
<sup>7</sup>L4 – fourth and final larval stage, <sup>8</sup>OECD - Organisation for Economic Co-operation and
Development, <sup>9</sup>UPLC-MS/MS - Ultra performance liquid chromatography - tandem mass
spectrometer, <sup>10</sup>WHOPES – World Health Organisation Pesticide Scheme

## 58 **1. Introduction**

Many studies have documented how the toxicity of pesticides such as organophosphates and 59 carbamates increases at higher mean temperatures (reviewed in Moe et al., 2013; Noyes et al., 60 61 2009; Noyes & Lema, 2015). In contrast, another key aspect of global warming, the increase in daily temperature variation (DTV; Colinet et al., 2015), has received much less attention in 62 ecotoxicology. The very few studies on this topic, however, convincingly showed that increased 63 DTV around the same mean temperature may, just as increases in mean temperature, magnify the 64 toxicity of pollutants (the pesticide chlorothalonil: Willming, Qin, & Maul, 2013; the 65 66 pharmaceutical fluoxetine: Barbosa et al., 2017).

Despite the growing interest in pesticide mixtures (e.g. Cedergreen, 2014; Hasenbein,
Lawler, & Connon, 2017; Hernández, Gil, & Lacasaña, 2017; Nowell et al., 2018), it is largely
unknown how thermal regimes shape their toxicity. One notable exception is the study of Laetz
et al. (2014), that documented that the synergistic effect in a mixture of two organophosphates
(ethoprop and malathion) was stronger at a higher mean temperature in the coho salmon

Oncorhynchus kisutch. Yet, no studies tested how DTV affects how pesticides within a mixture 72 interact and thereby determines mixture toxicity. When combined, pesticides may interact in 73 74 three ways (Cedergreen, 2014; Gunderson et al., 2016): (i) additive, where the combined effect equals the sum of the effects of the individual pesticides, (ii) synergistic, where the effect is 75 larger than the additive effect and (iii) antagonistic, where the effect is smaller than the additive 76 77 effect. The emerging view is that both synergistic and antagonistic effects in binary mixtures are rare and mostly occur at high concentrations (Cedergreen, 2014; Kamo and Yokomizo, 2015; 78 79 Rizzati et al., 2016; Rodea-Palomares et al., 2015). Evaluating the presence of synergistic effects between stressors is of key importance for nature management and risk assessment (Côté et al., 80 2016; Holmstrup et al., 2010). Given that synergisms can be expected when energetically costly 81 stressors are combined (Liess et al., 2016), and that DTV imposes net energetic costs as animals 82 face higher temperatures during their daily thermal cycle (Colinet et al., 2015), one can expect 83 DTV to increase the toxicity of pesticide mixtures. 84

85 While effects of thermal regimes on the toxicity of single pesticides and their mixtures are of large concern for populations of non-target organisms (Dinh Van et al., 2014; Relyea, 86 2009), they may at the same time improve the efficacy of the control of targeted pest species 87 88 (e.g. Gomes, Nunomura, Marinotti, & Tadei, 2016). Given the build-up of resistance, vector control strategies that rely on the use of single chemical insecticides are becoming increasingly 89 90 inefficient (Ranson and Lissenden, 2016), and this despite the increasing need for vector control 91 under global warming (Ogden and Lindsay, 2016). Integrated vector management (IVM) 92 promotes simultaneous use of different vector control tools and is therefore advocated to 93 improve the efficacy, cost-effectiveness, sustainability and ecological soundness of disease-94 vector control to prevent transmission of vector-borne diseases (WHO, 2012). The efficacy of

IVM based on pesticide mixtures (Bourguet et al., 2013) may obviously benefit from synergistic
interactions between the pesticides in the mixture used.

97 Synergistic interactions between biopesticides and chemical pesticides are a particularly promising IVM approach to control vector mosquitoes as these could potentially result in the 98 combination of efficacy of control, slowdown of resistance build-up and lower ecological 99 100 damage. One such promising biopesticide is Bti, the bacterium *Bacillus thuringiensis* var. israelensis (Becker et al., 2010). Bti is specifically harmful to dipterans and is the most widely 101 102 used biopesticide to control mosquitoes in Europe (Becker et al., 2010; Bellini et al., 2014; 103 WHOPES, 2016). The review of Cedergreen (2014) revealed that synergistic pesticide mixtures, while rare in general, often included an organophosphate. In line with this, a synergism was 104 detected when Bt or Bti were combined with the organophosphate temphos to kill Aedes aegypti 105 106 and Anopheles stephensi mosquitoes (de Andrande and Modolo, 1991; Narkhede et al., 2017). 107 However, an additive effect between Bti and the organophosphate malathion was found in the 108 mosquito *Culex pipiens* (Mansour et al., 2012). How DTV affects the toxicity of biopesticides has never been studied. 109

In this study, we tested whether DTV magnifies the toxicity of two pesticides with a 110 111 different mode of action, the organophosphate pesticide chlorpyrifos (CPF) and the biopesticide Bti, and of their mixture in a vector mosquito. Therefore, we setup up a full factorial design: 112 113 3 DTV levels (constant, low, high)  $\times$  2 CPF treatments (absent, present)  $\times$  2 Bti treatments 114 (absent, present). We expected that the toxic effects of the single pesticides are strengthened in 115 the presence of DTV (Barbosa et al., 2017; Willming et al., 2013). In addition, we tested whether there is an interaction between the chemical pesticide CPF and the biopesticide Bti within the 116 117 mixture, and whether this interaction is magnified in the presence of DTV. As study species we

used the Northern house mosquito *Culex pipiens* (Linaeus, 1758) form *molestus*. This species is
part of the *C. pipiens* complex, which are important vectors of the West Nile and St. Louis
encephalitis viruses (Farajollahi et al., 2011). In Europe and in the USA, this species is the most
common mosquito vector species in urban areas (Fonseca et al., 2004) which makes it an
important target for vector control (Kilpatrick, 2011). As chemical pesticide, we chose CPF as
this is one of the most used chemical insecticides in vector control management (Eaton et al.,
2008; WHOPES, 2016).

125 **2.** Material and methods

## 126 2.1 Study species and rearing

127 A continuous lab culture of C. pipiens initiated from a culture maintained at the Helmholtz 128 Centre for Environmental Research (UFZ, Germany) was used in this experiment. The original field population was also situated in Germany where it encountered a mean summer water 129 temperature of ca. 20°C (Tran et al. 2016). The mosquitoes were reared from egg to third instar 130 (L3) larvae in white 2 L containers (18.0 x 13.3 x 12.1 cm<sup>3</sup>) filled with 1 L aerated tap water. 131 Containers were placed in a temperature-controlled room under standard conditions of 14:10 h 132 light:dark and a water temperature of 19.84 °C (STD: 1.00 °C). We kept ca. 100 larvae (from 2-3 133 egg clutches) in each 2 L container. Larvae were fed three times a week with 3 mL of a 20 g/L 134 mixture of Olvarit<sup>®</sup> 7 cereal flakes (46%), wheat germs (51%) and Supradyn<sup>®</sup> vitamins (3%) (Op 135 136 de Beeck et al., 2016). This equals 0.257 mg of food/day/larva which is ad libitum.

## 137 2.2 Experimental design

138 To test the single and combined effects of daily temperature variation (DTV) and pesticide

139 exposure we used a full factorial design with three DTV treatments with the same mean

140	temperature of 20 °C (DTV-0 = constant temperature, DTV-7 = low daily temperature variation
141	of 7 °C, and DTV-14 = high daily temperature variation of 14 °C) crossed with two CPF
142	treatments (CPF absent and present) and with two Bti treatments (Bti absent and present). This
143	gave a full factorial 3 DTV $\times$ 2 CPF $\times$ 2 Bti experimental design. Note that the treatment
144	combination where both CPF and Bti are present, is the mixture treatment. Note also that our aim
145	was not to test for an effect of mean temperature but instead to test for an effect of DTV at a
146	given mean temperature. Therefore, and to keep the experiment feasible, we ran all DTV
147	treatments at the same mean temperature (see e.g. also Alto et al., 2017; Carrington et al., 2013;
148	Willming et al., 2013), here 20 °C the mean summer water temperature at the site of origin of the
149	study population. This design perfectly suits our aim to test the effect of realistic DTV scenarios.
150	We started 27-36 vials per treatment combination (total of 377 vials and 7,540 mosquito
151	larvae). The exact number of replicates of each treatment combination per response variable is
152	indicated above the bars in the figures. The experiment was performed in the last larval instar
153	(L4) during which larvae were continuously exposed to a DTV treatment, while the pesticide
154	exposure treatment was imposed for only two days. This resulted in three experimental steps:
155	(i) 4-day exposure to DTV, (ii) 2-day exposure to DTV and the pesticide treatment and
156	(iii) exposure to DTV until metamorphosis (Figure 1).
157	We started the DTV treatment within 24h after larvae molted into the last larval instar
158	(L4). At that moment, larvae were placed per 20 in 210 mL glass vials filled with 100 mL pre-
159	aerated water. Vials with larvae were placed in incubators at a 14:10 h light:dark photoperiod
160	with a light intensity of 1730 lux ( $\pm$ SE: 132 lux, $N = 4$ ) measured with a Testo 0500 Lux-meter.
161	All treatments received the same photoperiod and light intensity, and also experienced the same

162 mean temperature, only the daily temperature cycle differed between the three DTV treatments.

After four days, we quantified mortality and started the pesticide treatment. Per DTV treatment, 163 vials were randomly attributed to one of the four pesticide treatment combinations. At that 164 moment, we refreshed the medium with 100 mL of one of the four pesticide solutions. Two days 165 later, mortality was quantified again and the survivors were transferred to a new 210 mL glass 166 vial filled with 100 mL pre-aerated water while they stayed on their DTV treatment. We daily 167 168 counted larvae and adjusted food rations to keep the food level constant per larva (at 0.313 mg of 169 food/day/larva, Beketov & Liess, 2007); except during the two-day pesticide exposure period 170 when they did not receive any food following OECD guidelines (OECD, 2011).

171 After the 2-day pesticide exposure, we kept monitoring for delayed mortality, and recorded emergence success and development time (from egg hatching until metamorphosis). 172 Total mortality was calculated from the start of the L4 stage until metamorphosis. Successfully 173 174 emerged adults were sexed and frozen to measure the wing lengths of the females afterwards 175 with StreamPix v7 (Norpix, Inc., Canada) and ImageJ v1.51k (Rasband, W., National institute of 176 health, USA). Only the wing lengths of the females were measured, since this is a good proxy for their fecundity (Costanzo et al., 2011). Note that all end points (total mortality, development time 177 and adult wing length) are time-integrated 'accumulated' measures of the effects of the 178 179 treatments across the entire experiment, hence are not linked to a specific moment in the daily 180 temperature cycle.

#### 181 *2.3 DTV treatment*

At DTV-7 water temperatures fluctuated daily between 16.5 °C and 23.5 °C, while at DTV-14 water temperatures fluctuated between 13 °C and 27 °C. The water temperatures were measured every 15 minutes in a single vial per incubator during the experiment using HOBO temperature loggers. The water temperatures are shown in the appendix Figure A.1. The mean temperature of 20 °C in the two regimes with fluctuating temperatures (DTV-7 and DTV-14) was reached at
11h30 and 23h30; there was an increase in temperature from 8h30 to 20h30, and a decrease in
temperature from 20h30 to 8h30 in steps of 3 hours.

189 2.4 Pesticide treatment

Chlorpyrifos (CPF, purity grade > 99%) was purchased at Sigma-Aldrich (St. Louis, Missouri, 190 USA) and Bacillus thuringiensis israelensis (Bti) was bought as Vectobac WG Bacillus 191 192 thuringiensis serotype H-14 (37.4%) at Technic 3D (Paris, France). Based on a pilot study we 193 chose for both pesticides a concentration that caused moderate (ca. 30%) mortality, hence that 194 still allowed to detect sublethal effects: 70  $\mu$ g/L for Bti and 0.65  $\mu$ g/L for CPF. Both concentrations are ecologically relevant. According to WHOPES (2016) guidelines to control 195 196 mosquitoes, the recommended application dosage for natural water bodies ranges from 12.5 to 197 75 mg/m<sup>2</sup> for Vectobac WG and from 1.1 to 2.5 mg/m<sup>2</sup> for CPF. If a depth of 0.5 m is assumed 198 for a natural body, which corresponds with the shallow ponds preferred by *Culex pipiens* (Becker 199 et al., 2010), this gives a recommended dosage of 25-150  $\mu$ g/L for Bti and of 2.2-5  $\mu$ g/L for 200 CPF. The CPF solution was prepared by using a stock solution of 100 µg/mL CPF dissolved in 201 absolute ethanol which was kept in the dark at 4 °C. From this stock solution, a second stock 202 solution of 1 µg/mL was made in Milli-Q water. All pesticide solutions (including the Bti treatment) contained 104.5 µL ethanol/L. There were no differences in survival or growth rate of 203 L4 larvae of *C. pipiens* between the solvent control and the water control (results not shown). 204 205 The Bti solution was prepared daily from a new stock solution of 1 mg/mL Vectobac dissolved in Milli-Q water. From this stock solution, a second stock solution of 100 µg/mL was made in 206 207 Milli-Q water.

208	During the 2-day pesticide exposure period we gave one pesticide pulse at the start when
209	refreshing the medium. We took water samples from three vials of the CPF treatments at the start
210	and at the end of the pesticide exposure period which were analyzed at KU Leuven using
211	UPLC-MS/MS with Triple Quadrupole Mass Spectrometry. The initial CPF concentration was
212	$0.456 \pm 0.006 \mu$ g/L (mean ± SE, N = 3). After two days, the CPF concentration was 0.148 ±
213	0.013 $\mu$ g/L at DTV-0, 0.143 $\pm$ 0.011 $\mu$ g/L at DTV-7 and 0.099 $\pm$ 0.008 $\mu$ g/L at DTV-14 ( $N$ = 3).
214	This suggests that, if anything, CPF degradation across the 2-day exposure period was slightly
215	higher at DTV-14.
216	2.5 Statistical analyses
217	All statistical analyses were performed in R v3.3.2 (Core Team R, 2017) with the packages lme4
218	v1.1-13 (Bates et al., 2015), car v2.1-5 (Fox and Weisberg, 2002), afex v0.18-0 (Singmann et al.,
219	2017) and Ismeans v2.26-3 (Lenth, 2016). The figure was made in Microsoft Excel 2016.
220	Total mortality from the start of the L4 stage until metamorphosis was scored as 0 (alive)
221	and 1 (dead) for each larva within a vial. We tested for effects of the DTV (daily temperature
222	variation), CPF (chlorpyrifos) and Bti treatments on total mortality using a generalized linear
223	mixed model with a binomial error structure and the logit link. To test if the DTV, CPF and Bti
224	treatments had an effect on female wing length and female development time, we used general
225	linear mixed models with a normal error structure and the identity link. All models included all
226	possible interactions between the treatment factors. We only report the results on female wing
227	length (proxy for fecundity, Costanzo et al., 2011) and female development time as females are
228	the most important sex for population dynamics (Livdahl and Sugihara, 1984). For total
229	mortality, wing length and development time we used individuals as the unit of replication, yet

took into account that animals from the same vial were not independent by adding vial to the

models as random factor. If there was a significant effect of the DTV treatment (which had three
levels) or a significant interaction between the DTV, CPF and/or Bti treatments, we performed
Tukey HSD post hoc tests to further explore which DTV levels differed.

To identify which type of interaction was present between CPF and Bti in the mixture, the 234 independent action model was used as advocated when both chemicals have a different mode of 235 236 action and when analyzing mortality (Bliss, 1939; Faust et al., 2000; Schäfer and Piggott, 2018). Given that DTV may change the interaction type we determined this type per DTV level. We 237 followed the procedure explained in detail in Coors & De Meester (2008). First, the observed 238 effects of the single pesticides in absolute units were transformed to proportional effects using: 239  $E_i = \frac{(e_i - e_{control})}{(e_{max} - e_{control})}$  where  $e_i$  is the single-pesticide effect in absolute units,  $e_{control}$  is the effect of 240 the pesticide-free control in absolute units and e<sub>max</sub> is the maximum possible effect of that single 241 pesticide. For total mortality 100% was used as e<sub>max</sub>, while for the female development time 242 243 (46 days) and the female wing length (4.279 mm) the maximum values in the dataset were used as  $e_{max}$ . Second, the predicted joint effects were calculated using the formula  $E_{mix} = 1 - 1$ 244  $\prod^{i}$  (1 –  $E_{i}$ ). Third, the predicted joint effect  $E_{mix}$  was transformed back to absolute units to 245 allow comparison with the observed joint effects using the formula  $e_{mix} = E_{mix} *$ 246  $(e_{max} - e_{control}) + e_{control}$ . When the predicted joint effect was not in the 95% confidence 247 interval of the observed joint effect in absolute units, this was taken as evidence for a significant 248 deviation of observed from predicted joint effects. A significantly stronger observed joint effect 249 compared to the predicted joint effect indicated a synergistic interaction, while a significantly 250 251 weaker observed joint indicated an antagonistic interaction.

#### 252 **3. Results**

#### 253 *3.1 Total mortality*

- 254 Both single CPF exposure and single Bti exposure increased the total mortality from the start of
- the L4 stage until metamorphosis (Figure 2A, Table 1). There was a significant 3-way DTV  $\times$
- 256 CPF × Bti interaction (Figure 2A, Table 1): DTV did not increase mortality in the solvent control
- 257 (Tukey: all P > 0.74), in the single Bti exposure (all P > 0.67) and in the single CPF exposure
- (all P > 0.24), but DTV-14 increased mortality in the CPF-Bti mixture (all P < 0.001). According
- to the independent action model the CPF-Bti mixture was additive at DTV-0 and DTV-7, while it
- 260 was synergistic at DTV-14 (Figure 2A, Table 2).

## 261 *3.2 Development time of females*

- 262 Both single CPF exposure and single Bti exposure reduced the development time of females
- (Figure 2B, Table 1). There was a significant 3-way DTV  $\times$  CPF  $\times$  Bti interaction (Figure 2B,
- Table 1): DTV-14 decreased the development time in the solvent control (Tukey: all  $P \le 0.001$ ),
- in the single Bti exposure (all  $P \le 0.001$ ) and in the CPF-Bti mixture (all  $P \le 0.0.39$ ), but not in
- the single CPF exposure (P > 0.12). According to the independent action model the CPF-Bti
- 267 mixture was antagonistic at DTV-7, while it was additive at DTV-0 and DTV-14 (Figure 2B,
- 268 Table 2).

#### 269 *3.3 Wing length of females*

- 270 Single CPF exposure but not single Bti exposure increased the female wing length (Figure 2C,
- Table 1). There was a significant 3-way  $DTV \times CPF \times Bti$  interaction (Figure 2C, Table 1). This
- indicated that DTV did not affect the female wing length in the single Bti exposure (Tukey: all
- 273 P > 0.16) and in the single CPF exposure (all P > 0.11), but DTV-14 decreased wing length in

the solvent control (all  $P \le 0.023$ ) and increased the wing length in the CPF-Bti mixture (P = 0.018). According to the independent action model the CPF-Bti mixture was antagonistic for female wing length at DTV-7, while it was additive at DTV-0 and DTV-14 (Figure 2C, Table 2).

#### 278 **4. Discussion**

As expected, single exposure to each pesticide caused mortality and shaped the sublethal traits in
the survivors. Unexpectedly, daily temperature variation (DTV) did not make the single
pesticides more toxic. Yet, a key finding was that DTV did increase the toxicity of the mixture
by changing the interaction type between both pesticides for mortality from additive to
synergistic.

## 284 *4.1 Effects of the pesticides and their mixture at the constant temperature*

As expected, exposure to 0.65  $\mu$ g/L of the chemical pesticide CPF or to 70  $\mu$ g/L of the 285 286 biopesticide Bti increased the mortality of C. pipiens. In addition, both pesticides also reduced the female development time. Similarly, CPF exposure reduced the development time in the 287 damselfly *Enallagma cyathigerum* when enough food was available (Janssens & Stoks, 2013). 288 289 Earlier metamorphosis when exposed to pesticides in the larval stage may be an adaptive strategy 290 to escape the pesticide exposure in the water. Despite that a shorter development time gives less time to acquire resources (Dmitriew, 2011), the female wing length increased when exposed to 291 CPF. The beneficial sublethal effects of CPF may be explained by survival selection retaining 292 only survivors of high quality. In line with our results, females of the mosquito Aedes aegypti 293 294 that survived exposure to the insecticide Spinosad also had a larger wing length (Antonio et al., 2009). 295

Consistent with the emerging view on mixture toxicity (Cedergreen, 2014; Kamo and 296 Yokomizo, 2015; Rizzati et al., 2016; Rodea-Palomares et al., 2015), no endpoint showed a 297 298 synergistic interaction between CPF and Bti in their binary mixture when the temperature was kept constant (DTV-0°C). On the other hand, in the few studies that did document synergistic 299 interactions in mixtures, organophosphates are overrepresented (Cedergreen, 2014; see also de 300 301 Andrande and Modolo, 1991 and Narkhede et al., 2017). Two main mechanisms behind synergisms between organophosphates and other pesticides have been suggested (Cedergreen, 302 303 2014): (i) a change in the uptake rates and transport to the target site where one pesticide causes 304 an increase in oxygen consumption due to metabolization leading to an increased respiration, which increases the uptake of the other pesticide, and (ii) a change in metabolic enzyme 305 activities where one pesticide can increase the activation to a more toxic metabolite of the other 306 pesticide or can decrease the detoxification by enzymes. For example, the mixture of the 307 organophosphate temephos and the biopesticide Bt increased the level of alkaline phosphatase 308 309 that increases the toxicity of Bt compared to the control and single Bt exposure (Narkhede et al., 2017). The here documented additive effects between CPF and Bti at the constant temperature, 310 however, does correspond with the additive effect between another organophosphate, malathion, 311 312 and Bti for mortality in Culex pipiens (Mansour et al., 2012).

## 313 4.2 Effects of daily temperature variation (DTV) in the pesticide-free control

314 The imposed levels of DTV did not affect the mortality in the pesticide-free control. This

indicates that at a mean of 20 °C even a DTV of 14 °C was in general not harmful. At DTV-14,

- temperatures up to 27 °C are reached for 3 hours during each 24h cycle. While constant
- exposures to 24 °C have been shown to increase mortality in the study species (Tran et al.,
- 2016), it is known that under DTV animals may recover during the colder periods of each cycle

and therefore are able to endure short periods of otherwise harmful temperatures (Colinet et al., 319 2015). In line with this, a DTV of 12 °C did not increase total mortality around a mean of 26 °C 320 but did so around a mean temperature of 32 °C in the mosquito Anopheles stephensi (Paaijmans 321 et al., 2013). Similarly, a DTV of 18 °C around a mean of 26 °C caused an increase in larval 322 mortality in the mosquito Aedes aegypti (Carrington et al., 2013). 323 324 While the low DTV of 7 °C also had no sublethal effects on *C. pipiens*, the high DTV of 14 °C affected development time and wing length. The high DTV decreased the female 325 326 development time suggesting that at a mean of 20 °C the thermal performance curve for 327 development rate is concave upward. This indeed would imply that the increases in development rate at temperatures higher than the mean are larger than the decreases at temperatures lower 328 329 than the mean, resulting in a net increase of development rate under DTV (Colinet et al., 2015). Similarly, in A. stephensi a DTV of 12 °C around a low mean temperature (16 and 18 °C) also 330 resulted in shorter development times (Paaijmans et al., 2010). In contrast, a high DTV of 18 °C 331 332 caused an increase in development time of the mosquito A. *aegypti* (Carrington et al., 2013). However, the latter was around a higher mean temperature than in our study (26 °C compared to 333 20 °C). Furthermore, the high DTV of 14 °C reduced female wing length, matching the general 334 335 pattern of reductions in body size at higher DTV (Colinet et al., 2015). Given that the wing length is a good proxy for fecundity in C. pipiens (Costanzo et al., 2011), this indicates that a 336 337 large DTV is expected to decrease fecundity. This may be explained by the energetic costs of 338 coping with the daily experienced stressful upper temperatures associated with the high DTV 339 (Colinet et al., 2015).

## 340 *4.3 Effects of DTV on the toxicity of pesticides and their mixture*

Our expectation of DTV magnifying the toxic effects of single exposure to CPF or Bti was not 341 342 upheld. This expectation was based on the fact that during the daily thermal cycle higher temperatures are reached under DTV, and both CPF (e.g. Dinh Van et al., 2014; Hooper et al., 343 2013) and Bti (e.g. Duque and Navarro-Silva, 2006; Stevens et al., 2004) have been shown to be 344 345 more toxic at higher temperatures. Several mechanisms have been suggested for the higher toxicity of pollutants at higher temperatures, including an increase in uptake and metabolic 346 347 activity (Noyes et al., 2009). Synergistic interactions between chemical and environmental stressors are in general expected when there are energetic costs when coping with each single 348 stressor (Liess et al., 2016). It can therefore also be expected that DTV magnifies the effect of a 349 pollutant because DTV may generate energetic costs when organisms deal with the higher 350 temperatures during the cycle (Colinet et al., 2015). Accordingly, the study of Willming et al. 351 (2013) reported an increased mortality imposed by the chemical pesticides chlorothalonil (in 352 353 Hyalella azteca) and bifenthrin (in Chironomus dilutes) under DTV. Yet, in line with our findings, a DTV of 7.5 °C did not magnify the negative effect of the pharmaceutical fluoxetine 354 on survival in D. magna (Barbosa et al., 2017). The latter was explained by prioritization of 355 356 survival over reproduction in the presence of the combined stressors (DTV and CPF), as DTV did magnify the negative effect of fluoxetine on reproduction (Barbosa et al., 2017). While the 357 358 high DTV likely also caused energetic costs in our study (as suggested by the reduced wing 359 length), the costs may have been too low for DTV to magnify the effect of Bti alone or of CPF 360 alone.

A key novel finding of our study was that the high DTV did increase the mortalityimposed by the pesticide mixture which was associated with a change in the interaction type

between both pesticides in the mixture from additive at a constant temperature into synergistic at 363 high DTV. Notably, this occurred despite that the high DTV did not increase the toxicity of one 364 365 of the individual pesticides. Note that, given the CPF concentrations at the end of the experiment were, if anything, lower at high DTV (possibly due to a higher photodegradation), our finding of 366 a higher toxicity of the mixture is conservative. The specific mechanisms behind the interactions 367 368 between daily temperature variation and pesticides are not understood (e.g. Barbosa et al., 2017; 369 Willming et al., 2013). The energetic costs associated with coping with the higher temperatures 370 during the daily cycle (Colinet et al., 2015) may have only generated a synergism when energy 371 levels of the larvae are very low as can be expected in the mixture when larvae had to defend against both pesticides. Energetic costs may also explain why at a higher mean temperature the 372 synergism between two organophosphates was stronger in the coho salmon Oncorhynchus 373 kisutch (Laetz et al., 2014). A general explanation for why combinations of toxicants and 374 375 environmental stressors (such as DTV) may work synergistically is provided by the Stress 376 Addition Model (Liess et al., 2016). According to this model, each organism has the capacity to handle a certain level of stress. When stress levels become too high (as happens when stressors 377 378 are combined), individuals with a stress capacity below the given total stress level will die. The 379 SAM shows theoretically how a mechanism where both stressors jointly reduce the common stress capacity of organisms, translates into a synergistic interaction for mortality. This stress 380 381 capacity can be seen as the energy-limited tolerance to stress (sensu Sokolova, 2013). Thus, 382 environmental stress (in our case a high DTV), by being energetically costly (Colinet et al., 383 2015), can increase the toxicant sensitivity of individuals thereby creating a synergistic 384 interaction effect between CPF and Bti in their mixture at high DTV (Liess et al., 2016). In 385 addition, it may also be possible that under DTV, specifically at the higher temperatures during

the daily cycle, there was a higher uptake of both pesticides contributing to their synergism. 386 While our study was not designed to explain the type of interaction, this dependence of the 387 388 interaction type between two pesticides on DTV is intriguing and asks for mechanistic follow-up studies at the physiological level. Based on our hypothesized mechanism, particularly measuring 389 levels of energy storage and energy consumption may be rewarding. Whatever the underlying 390 391 reason, the appearance of this synergistic interaction between CPF and Bti only at high DTV 392 might suggest that interaction effects in binary mixtures are not so rare as based on meta-393 analyses (e.g. Cedergreen, 2014). Indeed, while daily temperature fluctuations are common in 394 nature, no other mixture studies considered natural daily temperature variation in their experimental design. 395

Intriguingly, DTV had a different effect on the interaction type between both pesticides 396 for the sublethal response variables. Indeed, it caused an antagonistic interaction for both 397 398 development time and wing length. In general, differences in interaction effects between the 399 same stressors for different end points are widespread, yet poorly understood (Côté et al., 2016). Development time decreased and wing length increased when exposed to the CPF-Bti mixture at 400 a constant temperature which corresponded to the sum of the response to the single pesticide 401 402 exposures (additive effect). However, these energetically costly responses apparently were not possible anymore to the same extent in the presence of the DTV stressor, likely because of a 403 404 higher total energetic cost. This might explain the antagonistic effect between CPF and Bti at a 405 low DTV for both sublethal endpoints. Noteworthy, at high DTV the interaction type became 406 additive again. This may be explained by survival selection associated with the synergistic effect 407 between CPF and Bti at high DTV for total mortality. Indeed the higher mortality in the mixture 408 at high DTV may have removed the weakest individuals with lower energy budgets.

Despite the increasing concern of warming magnifying the toxicity of single pesticides (Hooper 410 411 et al., 2013; Moe et al., 2013; Noyes and Lema, 2015), and the increasing numbers of studies on 412 the toxicity of pesticide mixtures (Cedergreen, 2014; Hernández et al., 2017), how temperature influences the interaction between two pesticides in a mixture has been largely ignored. The very 413 414 limited evidence shows that both a higher mean temperature (Laetz et al., 2014) and a higher daily temperature variation (this study) may increase the toxicity of pesticide mixtures. The 415 416 finding that DTV can modify the interaction effect between pesticides in a mixture is important 417 for vector control and ecological risk assessment. It indicates that DTV may not only change life history traits of vector species (Paaijmans et al., 2010; this study), but also the efficiency of 418 419 pesticide mixtures used in vector control. Our data suggest that laboratory-based individual 420 toxicity of chlorpyrifos or Bti obtained at constant temperatures may not be affected by the DTV 421 that is present under natural conditions, yet their mixture toxicity may become higher under 422 DTV. Applying realistic levels of DTV may therefore be important when designing laboratory 423 tests for developing pest management strategies based on pesticide mixtures. Future studies are 424 needed to test the generality of our findings across a broader range of concentrations and for 425 other mixtures, and especially for vector control multigenerational tests will be very valuable. Furthermore, mechanistic studies unraveling the mechanisms causing the interaction effects in 426 the presence of DTV will further enable making generalizations. The higher toxicity of the 427 428 mixture at high DTV compared to the typically used constant test temperatures in the laboratory 429 urges caution when evaluating both the vector control efficiency and the environmental impact 430 of pesticide mixtures as in nature daily temperature variation is omnipresent.

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439	7. References					
440	Alto, B.W., Wiggins, K., Eastmond, B., Ortiz, S., Zirbel, K., Lounibos, L.P., 2018. Diurnal					
441	temperature range and Chikungunya virus infection in invasive mosquito vectors. J. Med.					
442	Entomol. 55, 217–224. https://doi.org/10.1093/jme/tjx182					
443	Antonio, G.E., Sánchez, D., Williams, T., Marina, C.F., 2009. Paradoxical effects of sublethal					
444	exposure to the naturally derived insecticide spinosad in the dengue vector mosquito, Aedes					
445	aegypti. Pest Manag. Sci. 65, 323-326. https://doi.org/10.1002/ps.1683					
446	Barbosa, M., Inocentes, N., Soares, A.M.V.M., Oliveira, M., 2017. Synergy effects of fluoxetine					
447	and variability in temperature lead to proportionally greater fitness costs in Daphnia: A					
448	multigenerational test. Aquat. Toxicol. 193, 268–275.					
449	https://doi.org/10.1016/j.aquatox.2017.10.017					
450	Bates, D., Mächler, M., Bolker, B., Walker, S., 2015. Fitting linear mixed-effects models using					
451	lme4. J. Stat. Softw. 67, 1-48. https://doi.org/10.18637/jss.v067.i01					
452	Becker, N., Petrić, D., Zgomba, M., Boase, C., Madon, M., Dahl, C., Kaiser, A., 2010.					
453	Mosquitoes and their control, Second. ed. Springer, Verlag Berlin Heidelberg.					

- 455 Beketov, M.A., Liess, M., 2007. Predation risk perception and food scarcity induce alterations of
- 456 life-cycle traits of the mosquito *Culex pipiens*. Ecol. Entomol. 32, 405–410.
- 457 https://doi.org/10.1111/j.1365-2311.2007.00889.x
- 458 Bellini, R., Zeller, H., Van Bortel, W., 2014. A review of the vector management methods to
- 459 prevent and control outbreaks of West Nile virus infection and the challenge for Europe.
- 460 Parasit. Vectors 7, 323. https://doi.org/10.1186/1756-3305-7-323
- 461 Bliss, 1939. The toxicity of poisons applied jointly. Ann. Appl. Biol. 26, 585–615.
- 462 https://doi.org/https://doi.org/10.1111/j.1744-7348.1939.tb06990.x
- Bourguet, D., Delmotte, F., Franck, P., Guillemaud, T., Reboud, X., Vacher, C., Walker, A.-S.,
- 464 2013. Heterogeneity of selection and the evolution of resistance. Trends Ecol. Evol. 28,
- 465 110–118. https://doi.org/10.1016/j.tree.2012.09.001
- 466 Carrington, L.B., Seifert, S.N., Willits, N.H., Lambrechts, L., Scott, T.W., 2013. Large diurnal
- 467 temperature fluctuations negatively influence *Aedes aegypti* (Diptera: Culicidae) life-history
- 468 traits. Entomol. Soc. Am. 50, 43–51. https://doi.org/https://doi.org/10.1603/ME11242
- 469 Cedergreen, N., 2014. Quantifying synergy: A systematic review of mixture toxicity studies
- 470 within environmental toxicology. PLoS One 9, e96580.
- 471 https://doi.org/10.1371/journal.pone.0096580
- 472 Colinet, H., Sinclair, B.J., Vernon, P., Renault, D., 2015. Insects in fluctuating thermal
- 473 environments. Annu. Rev. Entomol. 60, 123–140. https://doi.org/10.1146/annurev-ento-
- 474 010814-021017
- 475 Coors, A. & De Meester, L., 2008. Synergistic, antagonistic and additive effects of multiple
- 476 stressors: predation threat, parasitism and pesticide exposure in *Daphnia magna*. J. Appl.

<sup>454</sup> https://doi.org/10.1007/978-3-540-92874-4

- 477 Ecol. 45, 1820–1828. https://doi.org/10.1111/j.1365-2664.2007.0
- 478 Core Team R, 2017. R: A language and environment for statistical computing. R Found. Stat.
- 479 Comput. https://doi.org/ISBN 3-900051-07-0
- 480 Costanzo, K.S., Muturi, E.J., Lampman, R.L., Alto, B.W., 2011. The effects of resource type and
- 481 ratio on competition with *Aedes albopictus* and *Culex pipiens* (Diptera: *Culicidae*). J. Med.
- 482 Entomol. 48, 29–38. https://doi.org/10.1603/ME10085
- 483 Côté, I.M., Darling, E.S., Brown, C.J., 2016. Interactions among ecosystem stressors and their
- 484 importance in conservation. Proc. R. Soc. B Biol. Sci. 283, 20152592.
- 485 https://doi.org/10.1098/rspb.2015.2592
- de Andrande, C.F.S., Modolo, M., 1991. Susceptibility of Aedes aegypti larvae to temephos and
- 487 *Bacillus thuringiensis var israelensis* in integrated control. Rev. Saude Publica 25, 184–
- 488 187. https://doi.org/10.1590/S0034-89101991000300004
- 489 Dinh Van, K., Janssens, L., Debecker, S., Stoks, R., 2014. Temperature- and latitude-specific
- 490 individual growth rates shape the vulnerability of damselfly larvae to a widespread
- 491 pesticide. J. Appl. Ecol. 51, 919–928. https://doi.org/10.1111/1365-2664.12269
- 492 Dmitriew, C.M., 2011. The evolution of growth trajectories: What limits growth rate? Biol. Rev.
- 493 86, 97–116. https://doi.org/10.1111/j.1469-185X.2010.00136.x
- 494 Duque, L.J.E., Navarro-Silva, M.A., 2006. Dynamics of the control of Aedes (Stegomyia) aegypti
- 495 Linnaeus (Diptera, *Culicidae*) by *Bacillus thuringiensis var israelensis*, related with
- 496 temperature, density .... Rev. Bras. Entomol. 50, 528–533. https://doi.org/10.1590/S0085-
- 497 56262006000400014
- 498 Eaton, D.L.L., Daroff, R.B.B., Autrup, H., Bridges, J., Buffler, P., Costa, L.G.G., Coyle, J.,
- 499 McKhann, G., Mobley, W.C.C., Nadel, L., Neubert, D., Schulte-Hermann, R., Spencer,

- 500 P.S.S., 2008. Review of the toxicology of chlorpyrifos with an emphasis on human
- 501 exposure and neurodevelopment. Crit. Rev. Toxicol. 38, 1–125.
- 502 https://doi.org/10.1080/10408440802272158
- 503 Farajollahi, A., Fonseca, D.M., Kramer, L.D., Kilpatrick, A.M., 2011. "Bird biting" mosquitoes
- and human disease: a review of the role of *Culex pipiens* complex mosquitoes in
- 505 epidemiology. Infect. Genet. Evol. 11, 1577–1585.
- 506 https://doi.org/10.1016/j.meegid.2011.08.013.
- 507 Faust, M., Altenburger, R., Backhaus, T., Bödeker, W., Scholze, M., Grimme, L.H., 2000.
- 508 Predictive assessment of the aquatic toxicity of multiple chemical mixtures. J. Environ.
- 509 Qual. 29, 1063. https://doi.org/10.2134/jeq2000.00472425002900040005x
- 510 Fonseca, D.M., Keyghobadi, N., Malcolm, C.A., Mehmet, C., Schaffner, F., Mogi, M., Fleischer,
- 511 R.C., Wilkerson, R.C., 2004. Emerging vectors in the *Culex pipiens* complex. Science. 303,
- 512 1535–1538. https://doi.org/10.1126/science.1094247
- 513 Fox, J., Weisberg, S., 2002. An {R} companion to applied regression, Second. ed, Sage
- 514 Publications. SAGE Publications, Inc, Thousand Oaks CA.
- 515 https://doi.org/10.1177/0049124105277200
- 516 Gomes, E.O., Nunomura, S.M., Marinotti, O., Tadei, W.P., 2016. Synergistic potential of
- 517 dillapiole combined with pyrethroids against mosquitoes. Vector Biol. J. 1, 8–11.
- 518 https://doi.org/10.4172/2473-4810.1000114
- 519 Gunderson, A.R., Armstrong, E.J., Stillman, J.H., 2016. Multiple stressors in a changing world:
- 520 the need for an improved perspective on physiological responses to the dynamic marine
- 521 environment. Ann. Rev. Mar. Sci. 8, 357–378. https://doi.org/10.1146/annurev-marine-
- 522 122414-033953

- Hasenbein, S., Lawler, S.P., Connon, R.E., 2017. An assessment of direct and indirect effects of
  two herbicides on aquatic communities. Environ. Toxicol. Chem. 36, 2234–2244.
- 525 https://doi.org/10.1002/etc.3740
- 526 Hernández, A.F., Gil, F., Lacasaña, M., 2017. Toxicological interactions of pesticide mixtures:
- 527 an update. Arch. Toxicol. 91, 3211–3223. https://doi.org/10.1007/s00204-017-2043-5
- 528 Holmstrup, M., Bindesbøl, A.M., Oostingh, G.J., Duschl, A., Scheil, V., Köhler, H.R., Loureiro,
- 529 S., Soares, A.M.V.M., Ferreira, A.L.G., Kienle, C., Gerhardt, A., Laskowski, R., Kramarz,
- 530 P.E., Bayley, M., Svendsen, C., Spurgeon, D.J., 2010. Interactions between effects of
- environmental chemicals and natural stressors: A review. Sci. Total Environ. 408, 3746–
- 532 3762. https://doi.org/10.1016/j.scitotenv.2009.10.067
- Hooper, M.J., Ankley, G.T., Cristol, D.A., Maryoung, L.A., Noyes, P.D., Pinkerton, K.E., 2013.
- 534 Interactions between chemical and climate stressors: A role for mechanistic toxicology in
- assessing climate change risks. Environ. Toxicol. Chem. 32, 32–48.
- 536 https://doi.org/10.1002/etc.2043
- Janssens, L., Stoks, R., 2013. Fitness effects of chlorpyrifos in the damselfly Enallagma
- 538 cyathigerum strongly depend upon temperature and food level and can bridge
- metamorphosis. PLoS One 8, e68107. https://doi.org/10.1371/journal.pone.0068107
- 540 Kamo, M., Yokomizo, H., 2015. Explanation of non-additive effects in mixtures of similar mode
- 541 of action chemicals. Toxicology 335, 20–26. https://doi.org/10.1016/j.tox.2015.06.008
- 542 Kilpatrick, A.M., 2011. Globalization, land use, and the invasion of West Nile virus. Science.
- 543 334, 323–327. https://doi.org/10.1126/science.1201010
- Laetz, C.A., Baldwin, D.H., Hebert, V.R., Stark, J.D., Scholz, N.L., 2014. Elevated temperatures
- 545 increase the toxicity of pesticide mixtures to juvenile coho salmon. Aquat. Toxicol. 146,

546 38–44. https://doi.org/10.1016/j.aquatox.2013.10.022

- Lenth, R. V., 2016. Least-squares means: The R package lsmeans. J. Stat. Softw. 69, 1–33.
  https://doi.org/10.18637/jss.v069.i01
- Liess, M., Foit, K., Knillmann, S., Schäfer, R.B., Liess, H.-D., 2016. Predicting the synergy of
- 550 multiple stress effects. Sci. Rep. 6, 32965. https://doi.org/10.1038/srep32965
- Livdahl, T.P., Sugihara, G., 1984. Non-linear interactions of populations and the importance of
  estimating per capita rates of change. J. Anim. Ecol. 53, 573–580.
- 553 https://doi.org/10.2307/4535
- 554 Mansour, S.A., Foda, M.S., Aly, A.R., 2012. Mosquitocidal activity of two Bacillus bacterial
- endotoxins combined with plant oils and conventional insecticides. Ind. Crops Prod. 35, 44–

556 52. https://doi.org/10.1016/j.indcrop.2011.06.001

- 557 Moe, S.J., De Schamphelaere, K., Clements, W.H., Sorensen, M.T., Van den Brink, P.J., Liess,
- 558 M., 2013. Combined and interactive effects of global climate change and toxicants on

populations and communities. Environ. Toxicol. Chem. 32, 49–61.

- 560 https://doi.org/10.1002/etc.2045
- 561 Narkhede, C.P., Patil, C.D., Suryawanshi, R.K., Koli, S.H., Mohite, B.V., Patil, S.V., 2017.
- 562 Synergistic effect of certain insecticides combined with *Bacillus thuringiensis* on mosquito
- 563 larvae. J. Entomol. Acarol. Res. 49, 6265. https://doi.org/10.4081/jear.2017.6265
- 564 Nowell, L.H., Moran, P.W., Schmidt, T.S., Norman, J.E., Nakagaki, N., Shoda, M.E., Mahler,
- 565 B.J., Van Metre, P.C., Stone, W.W., Sandstrom, M.W., Hladik, M.L., 2018. Complex
- 566 mixtures of dissolved pesticides show potential aquatic toxicity in a synoptic study of
- 567 Midwestern U.S. streams. Sci. Total Environ. 613–614, 1469–1488.
- 568 https://doi.org/10.1016/j.scitotenv.2017.06.156

- 569 Noyes, P.D., Lema, S.C., 2015. Forecasting the impacts of chemical pollution and climate
- 570 change interactions on the health of wildlife. Curr. Zool. 61, 669–689.
- 571 https://doi.org/https://doi.org/10.1093/czoolo/61.4.669
- 572 Noyes, McElwee, M.K., Miller, H.D., Clark, B.W., Van Tiem, L.A., Walcott, K.C., Erwin, K.N.,
- 573 Levin, E.D., 2009. The toxicology of climate change: Environmental contaminants in a
- 574 warming world. Environ. Int. 35, 971–986. https://doi.org/10.1016/j.envint.2009.02.006
- 575 OECD, 2011. OECD guideline for the testing of chemicals *Chironomus* sp., acute
- 576 immobilisation test. OECD 1–11. https://doi.org/10.1787/9789264069947-en
- 577 Ogden, N.H., Lindsay, L.R., 2016. Effects of climate and climate change on vectors and vector-
- 578 borne diseases: ticks are different. Trends Parasitol. 32, 646–656.
- 579 https://doi.org/10.1016/j.pt.2016.04.015
- 580 Op de Beeck, L., Janssens, L., Stoks, R., De Lin Op, B., Lizanne, J., Robby, S., 2016. Synthetic
- 581 predator cues impair immune function and make the biological pesticide Bti more lethal for
- vector mosquitoes. Ecol. Appl. 26, 355–366. https://doi.org/10.1890/15-0326
- 583 Paaijmans, K.P., Blanford, S., Bell, A.S., Blanford, J.I., Read, A.F., Thomas, M.B., 2010.
- 584 Influence of climate on malaria transmission depends on daily temperature variation. Proc.
- 585 Natl. Acad. Sci. 107, 15135–15139. https://doi.org/10.1073/pnas.1006422107
- 586 Paaijmans, K.P., Heinig, R.L., Seliga, R.A., Blanford, J.I., Blanford, S., Murdock, C.C., Thomas,
- 587 M.B., 2013. Temperature variation makes ectotherms more sensitive to climate change.
- 588 Glob. Chang. Biol. 19, 2373–2380. https://doi.org/10.1111/gcb.12240
- 589 Ranson, H., Lissenden, N., 2016. Insecticide resistance in African Anopheles mosquitoes: A
- 590 worsening situation that needs urgent action to maintain malaria control. Trends Parasitol.
- 591 32, 187–196. https://doi.org/10.1016/j.pt.2015.11.010

- 592 Relyea, R.A., 2009. A cocktail of contaminants: How mixtures of pesticides at low
- 593 concentrations affect aquatic communities. Oecologia 159, 363–376.
- 594 https://doi.org/10.1007/s00442-008-1213-9
- 595 Rizzati, V., Briand, O., Guillou, H., Gamet-Payrastre, L., 2016. Effects of pesticide mixtures in
- human and animal models: An update of the recent literature. Chem. Biol. Interact. 254,
- 597 231–246. https://doi.org/10.1016/j.cbi.2016.06.003
- 598 Rodea-Palomares, I., González-Pleiter, M., Martín-Betancor, K., Rosal, R., Fernández-Piñas, F.,
- 599 2015. Additivity and interactions in ecotoxicity of pollutant mixtures: some patterns,
- 600 conclusions, and open questions. Toxics 3, 342–369. https://doi.org/10.3390/toxics3040342
- 601 Schäfer, R.B., Piggott, J.J., 2018. Advancing understanding and prediction in multiple stressor
- research through a mechanistic basis for null models. Glob. Chang. Biol. 1–10.
- 603 https://doi.org/https://doi.org/10.1111/gcb.14073
- 604 Singmann, H., Bolker, B., Westfall, J., Aust, F., 2017. afex: Analysis of factorial experiments.
- 605 Sokolova, I.M., 2013. Energy-limited tolerance to stress as a conceptual framework to integrate
- the effects of multiple stressors. Integr. Comp. Biol. 53, 597–608.
- 607 https://doi.org/10.1093/icb/ict028
- 608 Stevens, M.M., Akhurst, R.J., Clifton, M.A., Hughes, P.A., 2004. Factors affecting the toxicity
- 609 of *Bacillus thuringiensis var. israelensis* and *Bacillus sphaericus* to fourth instar larvae of
- 610 *Chironomus tepperi* (Diptera: *Chironomidae*). J. Invertebr. Pathol. 86, 104–110.
- 611 https://doi.org/10.1016/j.jip.2004.04.002
- Tran, T.T., Janssens, L., Dinh, K. V., Op de Beeck, L., Stoks, R., 2016. Evolution determines
- 613 how global warming and pesticide exposure will shape predator-prey interactions with
- 614 vector mosquitoes. Evol. Appl. 9, 818–830. https://doi.org/10.1111/eva.12390

- 615 WHO, 2012. Handbook for integrated vector management. World Heal. Organ. 78.
- 616 https://doi.org/10.1564/v24\_jun\_14
- 617 WHOPES, 2016. Recommended compounds and formulations for control of mosquito larvae
- 618 [WWW Document]. World Heal. Organ. URL
- 619 http://www.who.int/whopes/Mosquito\_larvicides\_Feb\_2016.pdf
- 620 Willming, M.M., Qin, G., Maul, J.D., 2013. Effects of environmentally realistic daily
- 621 temperature variation on pesticide toxicity to aquatic invertebrates. Environ. Toxicol. Chem.
- 622 32, 2738–2745. https://doi.org/10.1002/etc.2354

Table 1. Results of the generalized linear mixed model (total mortality) and the general linear

mixed models (female development time and female wing length) testing for the effects of

626 exposure to chlorpyrifos (CPF) and to *Bacillus thuringiensis* var. *israelensis* (Bti) under different

627 levels of daily temperature variation (DTV) in *Culex pipiens*. *P*-values indicated in bold are

628 significant ( $P \le 0.05$ ).

	Total mortality		Female development time			Female wing length			
	$\chi^2$	df	Р	$\chi^2$	df	Р	$\chi^2$	df	Р
DTV	18.32	2	<0.001	57.36	2	<0.001	1.11	2	0.58
CPF	1449.32	1	<0.001	174.18	1	<0.001	79.86	1	<0.001
Bti	104.93	1	<0.001	42.84	1	<0.001	2.30	1	0.13
DTV×CPF	23.62	2	<0.001	5.05	2	0.080	13.40	2	0.001
DTV×Bti	9.67	2	0.008	1.90	2	0.39	6.41	2	0.041
CPF×Bti	0.51	1	0.47	1.10	1	0.30	0.02	1	0.90
DTV×CPF×Bti	10.43	2	0.005	6.53	2	0.038	5.99	2	0.050

629

Table 2. Results of the independent action models identifying the interaction type per DTV
(daily temperature variation) treatment for (A) the total mortality, (B) the female development
time, and (C) the female wing length. If the predicted value is not within the range of the 95%
confidence interval of the observed value, there is a synergism or an antagonism (indicated in
bold).

	Predicted	Observed	95%CI Observed	Interaction effect
(A) Total Mortality				
DTV of 0°C	84.49	83.33	[76.91, 89.76]	Additive
DTV of 7°C	86.09	85.96	[78.63, 93.30]	Additive
DTV of 14°C	86.79	94.19	[92.31, 96.07]	Synergistic
(B) Development Time				
DTV of 0°C	28.62	29.17	[28.38, 29.97]	Additive
DTV of 7°C	28.07	30.15	[29.09, 31.22]	Antagonistic
DTV of 14°C	26.90	27.00	[26.14, 27.86]	Additive
(C) Wing Length				
DTV of 0°C	3.64	3.70	[3.63, 3.76]	Additive
DTV of 7°C	3.75	3.66	[3.57, 3.74]	Antagonistic
DTV of 14°C	3.75	3.80	[3.71, 3.90]	Additive

## 637 Figure legends

Figure 1. Schematic overview of the experimental design.

639 Figure 2. Effects of chlorpyrifos (CPF), Bacillus thuringiensis var. israelensis (Bti) and the CPF-

- 640 Bti mixture under different levels of daily temperature variation (DTV) in *Culex pipiens*: (A)
- total mortality, (B) female development time, and (C) female wing length. Means are given with
- their 95% confidence interval. Numbers inside the bars indicate the number of vials (total
- 643 mortality) or number of adults (development time and wing length). Besides the observed values
- 644 for the CPF-Bti mixture also the predicted values of the mixture based on the independent action
- model are shown. An asterisk (\*) above the observed mixture values for a given DTV indicates a
- 646 significant difference (P < 0.05) with the predicted mixture values based on the independent
- 647 action model, hence a deviation from additivity. Interaction types for the CPF-Bti mixture per
- 648 DTV level are indicated as additive, synergistic or antagonistic.

# 650 Figure 1.

	20°C with DTV of 0°C (constant)					
	constant temperature of 20°C	↓ 20°C with D	DTV of 7°C			
Eggs	Last larval stage L4	↓ 20°C with D	Adults			
	clean water till last larval instar (L4) 🔌	2-day pesticide exposure - Solvent control - Chlorpyrifos (CPF) - Bacillus thuringiensis israelensis (Bti) - CPF-Bti mixture	clean water till emergence 💥			

