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How effective are soil conservation techniques in reducing plot runoff and soil loss in Europe and the Mediterranean?

W. Maetens^{*,1}, J. Poesen¹, M. Vanmaercke^{1,2}

Abstract

The effects of soil and water conservation techniques (SWCTs) on annual runoff (R_a) , runoff coefficients (RC_a) and annual soil loss (SL_a) at the plot scale have been extensively tested on field runoff plots in Europe and the Mediterranean. Nevertheless, a comprehensive overview of these effects and the factors controlling the effectiveness of SWCTs is lacking. Especially the effectiveness of SWCT in reducing R_a is poorly understood. Therefore, an extensive literature review is presented that compiles the results of 101 earlier studies. In each of these studies, R_a and SL_a was measured on field runoff plots where various SWCTs were tested. In total, 353 runoff plots (corresponding to 2093 plot-years of data) for 103 plot-measuring stations throughout Europe and the Mediterranean were considered. SWCTs include (1) crop and vegetation management (i.e. cover crops, mulching, grass buffer strips, strip cropping and exclosure), (2) soil management (i.e. no-tillage, reduced tillage, contour tillage, deep tillage, drainage and soil amendment) and (3) mechanical methods (i.e. terraces, contour bunds and geotextiles). Comparison of the frequency distributions of SL_a rates on cropland without and with the application of SWCTs shows that the exceedance probability of tolerable SL_a rates is ca. 20% lower when SWCT are applied. However, no notable effect of SWCTs on the frequency distribution of RC_a is observed. For 224 runoff plots (corresponding to 1567 plot-year data), SWCT effectiveness in reducing R_a and/or SL_a could be directly calculated by comparing measured R_a and/or SL_a with values measured on a reference plot with conventional management. Crop and vegetation management techniques (i.e. buffer strips, mulching and cover crops) and mechanical techniques (i.e. geotextiles, contour bunds and terraces) are generally more effective than soil management techniques (i.e. no-tillage, reduced tillage and contour tillage). Despite being generally less effective, no-tillage, reduced tillage and contour tillage have received substantially more attention in the literature than the other SWCTs. Soil and water conservation techniques are generally less effective in reducing R_a than in reducing SL_a , which is an important consideration in areas where water is a key resource and in regions susceptible to flooding. Furthermore, all SWCTs show a more consistent and effective reduction of both R_a and SL_a with increasing R_a and SL_a magnitude, which is attributed to the reduced influence of measurement uncertainties. Although some significantly negative correlations between SWCT effectiveness and plot slope length, slope gradient or annual precipitation were found, the importance of these factors in explaining the observed variability in effectiveness seems limited. Time-series analyses of R_a during multiple years of SWCT application strongly indicate that no-tillage and conservation tillage become less effective in reducing R_a over time. Such an effect is not observed for SL_a .

<u>Keywords:</u> soil and water conservation techniques, runoff reduction, soil loss reduction, crop and vegetation management, soil management, mechanical methods

¹ Division of Geography, Department of Earth and Environmental Sciences, KU Leuven, Celestijnenlaan 200E, B-3001 Heverlee, Belgium

² Research Foundation Flanders (FWO), Belgium

^{*} Corresponding author: willem.maetens@ees.kuleuven.be

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1 Introduction

Soil and Water Conservation Techniques (SWCTs) have long existed as a means to combat the detrimental effects of soil loss through interrill and rill erosion (Cerdà et al., 2009; Montgomery, 2007a; Morgan, 2005). The aim of SWCTs is to reduce both on-site runoff (R) and soil loss (SL) as well as the off-site consequences of erosion such as sedimentation of reservoirs, deterioration of water quality and flooding (e.g. Owens et al., 2005; Vanmaercke et al., 2011a; Verstraeten and Poesen, 1999). Recent research also focuses on the role of SWCTs in the conservation of various ecosystem functions of the soil and its role in bio-geochemical cycles, including carbon sequestration (e.g. Conley, 2000). Whereas the role of SWCTs in reducing soil loss is well recognised (e.g. Boardman and Poesen, 2006; Morgan, 2005), there is still a need to integrate SWCTs effectively into good agricultural and sustainable land management practices. This need is exemplified by the goals of recent policy developments such as the European Commission's Thematic Strategy for Soil Protection (European Commission, 2012). Furthermore, several international scientific projects focus on both quantifying the effectiveness of different SWCTs in reducing R and SL as well as on their successful implementation (e.g. DESIRE, 2007; Karlen, 2008; Louwagie et al., 2009; Römkens, 2010).

Successful SWCT application schemes are sufficiently effective in reducing R and SL to sustainable levels, while not being overdimensioned so that they are economically feasible. Implementing successful schemes therefore requires both qualitative assessments of the effects and applicability of SWCTs (e.g. Schwilch et al., 2011) as well as reliable quantitative data on of the R and SL reduction by the SWCT. The most widely used measure to quantify the effectiveness of SWCTs in reducing SL is the soil loss ratio (SLR), i.e. the ratio of SL from a plot with SWCT application and SL from a reference plot with the same characteristics but without SWCT application (e.g. Castillo et al., 1997; Cogo et al., 1984; Gilley and Risse, 2000; Smets et al., 2008a). SLR values are similar to the widely used (R)USLE cover management (C) and support practice (P) factor (Renard et al., 1997). However, the calculation of C- of P-factors for specific soil conservation techniques is not straightforward and the validity of the empirical relations for C- and P-factors given by Renard et al. (1997) outside the Midwestern U.S.A. is uncertain. Quantification of SWCT effectiveness for other regions requires local measurements of SLR (e.g. Hessel and Tenge, 2008).

Furthermore, C- and P-factors apply only to SL and not to R. While runoff ratios (RR), the equivalent of SLR, have been used in some studies (e.g. Gilley and Risse, 2000), quantification of SWCT effectiveness remains mainly oriented at SL. Nevertheless, the term 'soil and water conservation techniques' implies that also an effect on runoff is expected or desired. Despite the limited attention, runoff reduction remains an important concern. On-site, conservation of plant-available water is an important issue for agricultural production (Rockström et al., 2010; Wallace, 2000) and may be a more important concern than soil loss, e.g. in areas where water is a key resource. Furthermore, sediment yield at the catchment scale is in many cases strongly controlled by the occurence and magnitude of a few flood events (e.g. Gonzalez-Hidalgo et al., 2010). Hence, runoff reduction is a crucial part of integrated catchment management (Nyssen et al., 2010; Vanmaercke et al., 2010; Verstraeten and Poesen, 1999). In addition, runoff generation and soil loss for various land use types are closely related (Maetens et al., 2012) and information on the effectiveness of SWCTs in reducing R can also improve insights in their effectiveness in reducing SL.

There are also strong indications that the effectiveness of SWCT depends on environmental factors such as land use, saturated conductivity and storm size (Hessel and Tenge, 2008) or plot slope length (e.g. Gilley and Risse, 2000; Smets et al., 2008b,a) and plot slope gradient (e.g. Renard et al., 1997; Syversen, 2005). Nevertheless, very few quantitative assessments of the effects of these environmental factors on SWCT effectiveness in reducing R and SL have been made. Limited understanding of environmental effects on SWCT effectiveness in reducing R and SL also limits the incorporation of SWCT application in erosion models. (e.g. Hessel and Tenge, 2008). Finally, a comprehensive assessment of the effectiveness of SWCTs also needs to consider temporal aspects of SWCT application: the temporal variability in SWCT effectiveness and how this effectiveness evolves over the years since the initial application. The latter has been studied for the build-up of soil organic carbon (Hao et al., 2002), soil biochemical properties (Madejón et al., 2009) and crop yield (Rusinamhodzi et al., 2011; Van den Putte et al., 2010). However, no such study exists with respect to the long-term effects of SWCTs on R or SL.

An overview and meta-analysis of available field-measured data on the effectiveness of various SWCTs in reducing both R and SL can provide important additional insights and can improve our ability to model the effects of SWCT in reducing R and SL under various conditions. However, relatively few comprehensive overviews are currently available. A global assessment of SWCT effectiveness in reducing SL was made by Montgomery (2007b), but this analysis does not include R nor does it allow a quantification of the effectiveness of specific techniques. For Europe and the Mediterranean, available overviews of SWCTs effectiveness are very limited (Table 1) Several of the available overviews of erosion rates and their controlling factors do not consider SWCTs explicitly (Table 1). Furthermore, the reviews that consider SWCTs often do not include the effectiveness of SWCTs in reducing R. Studies that do consider R are limited to a few specific techniques (Table 1).

Table 1: Overview of compilation studies reporting annual runoff and soil loss data from experimental plots, including both conventional practices and the application of soil and water conservation techniques (SWCTs), collected at country, regional or subcontinental scale for Europe and the Mediterranean. R_a : annual runoff, SL_a : annual soil loss, SWCT: soil and water conservation technique, #PY= number of plot-years. +: reported in study, -: not reported in the study, C: cropland, V: vineyard, S, shrubland, T: tree crops, Pf: post-fire, Fo: forest, Ba: bare, Fa: fallow, Gr: grassland, NA: not available.

region/country	\mathbf{R}_a	\mathbf{SL}_a	SWCT	land use types	\mathbf{SWCT}	$\#\mathbf{PY}$	source
Mediterranean:							
MEDALUS project field sites; Spain, France, Italy, Greece	+	+	1	Cr, Vi, Tc, Sh, Eucalyptus	NA	NA	Kosmas et al., 1997
Portugal, Spain, France, Italy, Greece, Israel	1	+		Cr, Vi, Tc, Sh	NA	ca. 150	Wainwright and Thornes, 2004
Spain, Portugal, Greece, Syria	+	+ -	+ -	olive orchards	no-tillage, cover strips	45	Fleskens and Stroosnijder, 2007
Spain, France, Italy, Morocco		+	+	Or, Sil	rerraces	8C .	Golfzalez-fildalgo et al., 2007
Italy, Spain	ı	+	+	Sh, Pf, Fo, Cr, Ba, Fa, Gr	sludge application, burnt mulch, no-tillage, conservation tillage, contour tillage	74	de Vente, 2009; de Vente et al., 2007
Portugal, Spain, France, Italy, Greece, Israel, Croatia	ı	+	1	Pf, Cr, Vi, Sh, Eucalyptus	N.A.	NA	Shakesby, 2011
Germany	1	+	+	Cr, Fa, Gr, Vi, Fo	cover crop, mulch tillage, grass strips	1 078	Auerswald et al., 2009
Europe	1	+	+	Ba, Ba, Fo, Gr, Sh, Vi, Tc	cover crops	2 741	Cerdan et al., 2006, 2010
Europe	,	+	+	Cr	mulch	NA	Smets et al., 2008a
Europe	+	+	+	Cr	conservation tillage	NA	Leys et al., 2010
Europe		+	1	Ba, Cr, Fa, Fo, Gr, Pf, Sh, Tc, Vi	NA	NA	Boardman and Poesen, 2006

Therefore, the objectives of this paper are (1) to provide an overview of field plot data on effectiveness of SWCT in reducing annual runoff (R_a , $mm \cdot yr^{-1}$) and annual soil loss (SL_a , $Mg \cdot ha^{-1} \cdot yr^{-1}$) in Europe and the Mediterranean, (2) to quantify the effectiveness of different SWCT types in reducing both R_a and SL_a and to explore the effect of SWCTs on the relations between R_a and SL_a and (3) to explore the relations of SWCT effectiveness with some important variables that were reported in the experimental studies (i.e. magnitude of R_a and SL_a , plot length, plot slope gradient, annual precipitation (P_a , $mm \cdot yr^{-1}$), and the number of consecutive years of SWCT application).

2 Data collection

Annual runoff and soil loss data, measured on bounded plots where a SWCT was applied, for Europe and the Mediterranean were collected from research papers, books, project reports and PhD. theses (Figure 1). Each plot represents a combination of a soil type, a plot length, a slope gradient, and a land use type and is associated with one type of SWCT (Table 2). SWCTs were classified into three groups according to Morgan (2005) (Table 2). Only runoff and soil loss measurements from bounded runoff plots under natural rainfall, with a minimum length of 5m were retained. Only annual data are considered: either plot data were collected during at least a full year, or the reported data could be extrapolated to represent a full year with a sufficient degree of reliability, i.e. when measurements were conducted for at least 80% of the year and rainfall was uniformly distributed throughout the year (Maetens et al., 2012), or when authors indicated that the measurements were representative for a full year. For each plot, the corresponding number of plot-years (#PY) was determined, whereby 1 PY corresponds to a measuring period of 1 year on a single runoff plot.

Table 2: List of soil and water conservation techniques (SWCTs) tested on runoff and soil loss plots in Europe and the Mediterranean. Ba: bare soil, Cr: cropland, Cs: construction sites, Fa: fallow, Fo: forest, Gr: grassland, Pf: post-fire, Sh: shrubland, Tc: tree crops, Vi: vineyard

SWCT		Description
Crop and	vegetation management	
Cc:	cover crops	cover crop during the intercropping season, established after harvest (e.g. Laloy and Bielders, 2010).
Mu:	mulching	application of stone or organic mulch (crop residue or straw) (e.g. Brown, 1996).
Bs:	buffer strips	Strips of perennial vegetation (usually grasses) used to increase infiltration, slow down runoff and increase sediment deposition (e.g. Uusi-Kämppä, 2005).
Sc:	strip cropping	drilling or planting in strips of alternating crop types (e.g. Köse et al., 1996; Köse and Taysun, 2002).
Ex:	exclosure	closing of areas for grazing and agriculture (e.g. Bruggeman et al., 2005; Masri et al., 2005; Mazour, 1992).
Soil mana	gement	
Nt: Rt:	no-tillage reduced tillage	no tillage operations when the traditional practice is tillage (e.g. Turtola et al., 2007). different forms of reduced tillage resulting in a smaller disturbance of the plough layer than conventional cultivationKwaad et al. (1998).
Ct:	contour tillage	tillage operations parallel the contour (e.g. Quinton and Catt, 2004).
Dt:	deep tillage	deep non-inversion tillage to improve infiltration (e.g. Chomanicová, 1988; Suchanic, 1987).
Dr:	drainage	application of subsurface drainage pipes (e.g. Øygarden, 1996; Øygarden et al., 1997).
Sa:	soil amendment	application of soil conditioners or fertilizer to improve soil soil structural stability: i.e. phosphogypsum (e.g. Agassi and Benhur, 1991), polymers(e.g. Chomanicová, 1988; Lopez-Bermudez et al., 1991; Romero-Díaz et al., 1999), sewage sludge (e.g. Ingelmo et al., 1998) and NPK fertilizer (e.g. Kroumov and Malinov, 1989).
Mechanica	al methods	
Te:	terraces	construction of earthen or stone terraces parallel the contour (e.g. Koulouri and Giourga, 2007).
Cb:	contour bunds	stone or earthen bunds constructed parallel to the contour (e.g. Pinczés, 1982)
Gt:	geotextile	application of geotextile mats (e.g. Bhattacharyya et al., 2008, 2009; Jankauskas et al., 2008; Mitchell et al., 2003).

Most plots were equipped with tanks for collecting runoff and soil loss. However, for a small number of plots (n=27),soil loss was determined by measuring rill volumes (Feiza et al., 2007; Jankauskas and Jankauskiene, 2003; Jankauskas et al., 2007). These volumetric measurements were converted to SL_a data (in $Mg \cdot ha^{-1} \cdot yr^{-1}$) by assuming a soil bulk density of 1.5 g·cm⁻³ and

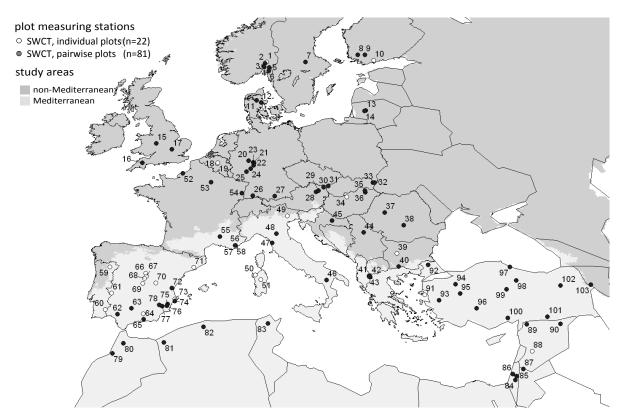


Figure 1: Geographical distribution of runoff and soil loss plot measuring stations included in the individual and paired plot SWCT databases for Europe and the Mediterranean. Black circles represent stations included in the paired plot database, while open circles represent stations which are only included in the individual plot database. The division between Mediterranean and Non-Mediterranean was derived from the LANMAP2 database (Mücher et al., 2010; Metzger et al., 2005). n= number of plot measuring stations.

by adding 25% to account for interrill erosion. This value was based on a literature review by Govers and Poesen (1988). If available, also the plot length, plot gradient, annual precipitation and measuring period were included in the data compilation.

3 Characteristics of the collected plot data

Figure 1 shows the SWCTs that have been tested in Europe and the Mediterranean. Relatively few plot data are available for Northern latitudes and for Eastern European countries compared to other regions. This may be partly explained by the fact that results from most plot studies in Eastern Europe are only reported in the local language and have not been published internationally. Nevertheless, this study represents the largest compilation of SWCT effectiveness measurements in Europe and the Mediterranean. An overview of the collected data per country is given in Table 3. Analyses of these data revealed no clear regional differences in the type of SWCTs evaluated (Table 3).

SWCTs have been tested on runoff plots in Europe and the Mediterranean since 1956 (Jung and Brechtel, 1980). Since then, the number of plots gradually increased and peaked between 1987 and 1997 (Figure 2a). This peak is considerably later than the majority of runoff plot measurements in the USA (between 1930 and 1942; Laflen and Moldenhauer, 2003) and the majority of catchment sediment yield studies in Europe (Vanmaercke et al., 2011b). The decline in the number of plots after 1997 is partly due to the termination of some large research projects (e.g. MEDALUS). Furthermore, recently collected plot data may not have been published yet. However, measuring runoff and soil loss from plots is labour- and cost-intensive and the decline

Table 3: Overview of the different soil and water conservation techniques (SWCTs) included in the global plot database and publications from which annual runoff and soil loss data were extracted in each country in Europe and the Mediterranean. References followed by * report data for plots where only SWCTs were applied, other references contain plots included in the paired plot database. References followed by TS were used for the analysis of plot data time-series (consecutive years) of SWCT application. Numbers between brackets refer to the plot measuring stations indicated in Figure 1.

Country	SWCT	Source
Albania	notillage reduced tillage drainage	Grazhdani et al., 1999 [42] Grazhdani et al., 1999 [42] Grazhdani et al., 1996 [41-43]; Grazhdani, pers. comm. [41-43]
Algeria	cover crops buffer strips exclosure notillage contour tillage	Arabi and Roose, 1993 [82] Arabi and Roose, 1993 [82] Mazour, 1992 [81]; Morsli et al., 2004 [81] Arabi and Roose, 1993 [82] Mazour, 1992 [81]; Mazour et al., 2008 ^{TS} [81]
Austria	cover crops	Klik, 2003 [28-30]; Klik, 2010; Klik, pers. comm.
Belgium	cover crops notillage reduced tillage	Laloy and Bielders, 2008*;Laloy and Bielders, 2010* [18,19] Laloy and Bielders, 2008*;Laloy and Bielders, 2010* [18,19] Laloy and Bielders, 2008*;Laloy and Bielders, 2010* [18,19]
Bulgaria	buffer strips soil amendment	Biolchev, 1975* [39]; Malinov, 1999* [39], cited by Rousseva et al., 2006 Kroumov and Malinov, 1989 [40]
Croatia	notillage contour tillage	Basic et al., 2001 , 2004^{TSTS} [45] Basic et al., 2001 , 2004^{TS} [45]
Denmark	cover crops reduced tillage	Schjønning et al., 1995* [11,12] Schjønning et al., 1995* [11,12]
Finland	cover crops buffer strips notillage reduced tillage contour tillage drainage	Turtola et al., 2007* [9] Puustinen et al., 2005, 2007 [8]; Uusi-Kämppä, 2005 [9] Puustinen et al., 2005, 2007 [8]; Turtola et al., 2007 TS [9] Puustinen et al., 2005, 2007 [8]; Turtola et al., 2007 TS [9] Puustinen et al., 2005, 2007 [8] Warsta et al., 2009* [10]
France	cover crops mulching buffer strips notillage reduced tillage	Messer, 1980 [54]; Viguier, 1993 [56-58] Ballif, 1989 [53]; Viguier, 1993 [56-58] AREDVI, 2003 [55]; Le Bissonnais et al., 2004 [52] Messer, 1980 [54]; Viguier, 1993 ^{TS} [56-58] Ballif, 1989 [53]
Germany	cover crops buffer strips reduced tillage contour tillage	Emde, 1992 [25]; Jung and Brechtel, 1980 [20-23,26-27] Jung and Brechtel, 1980* [20-23,26-27]; Voss, 1978 [24] Jung and Brechtel, 1980* [20-23,26-27] Jung and Brechtel, 1980* [20-23,26-27]
Greece	terraces	Koulouri and Giourga, 2007* [91]
Hungary	contour tillage contour bunds geotextile	Hudek and Rey, 2009* [34]; Pinczés, 1982 [36] Pinczés, 1982 [36] Kertesz et al., 2007 [35]; Kertesz, pers. comm. [35]
Israel	soil amendment	Agassi and Benhur, 1991 [86]; Agassi et al., 1990 [86]
Italy	cover crops mulching notillage reduced tillage drainage geotextile	Bini et al., 2006* [49]; Caredda et al., 1997* [50] Rivoira et al., 1989* [51] Postiglione et al., 1990 [46] Basso et al., 1983 ^{TS} [46]; Chisci and Zanchi, 1981* [47]; Chisci, 1989 [47,48]; Postiglione et al., 1990 [46] Chisci and Zanchi, 1981 [47] Zanchi, 1983 [48]
Jordan	mulching	Abu-Zreig, 2006 [87]; Abu-Zreig et al., 2011 [87]
Lithuania	cover crops notillage reduced tillage geotextile	Jankauskas and Jankauskiene, 2003* [13]; Jankauskas et al., 2007* [13] Feiza et al., 2007 [13] Feiza et al., 2007 [13] Jankauskas et al., 2008 [14]
Morocco	reduced tillage contour tillage	Heusch, 1970 [80] Laouina et al., 2003^{TS} [79]
Norway	reduced tillage drainage	Børresen, pers. comm. [34]; Grønsten and Lundekvam, 2006^{TS} [34]; Lundekvam, 2007^{TS} [2,5-6] Øygarden, 1996 [1]; Øygarden et al., 1997* [1]
Palestinian territories	terraces	Abu Hammad et al., 2004, 2006 [85]; Al-Seekh and Mohammad, 2009 [84]
Portugal	cover crops notillage contour tillage	Roxo et al., 1996* [60] Oliveira, 2005* [59] Oliveira, 2005* [59]
Romania	notillage reduced tillage terraces	Nistor and Ionita, 2002 [37] Nistor and Ionita, 2002 [37] Teodorescu and Badescu, 1988 [38]
Serbia	terraces	Djorović, 1990 [44]
Slovakia	mulching notillage reduced tillage deep tillage soil amendment	Chomanicová, 1988 [33] Fulajtár and Janský, 2001 [31] Chomanicová, 1988 [33]; Suchanic, 1987 [32] Chomanicová, 1988 [33]; Suchanic, 1987 [32] Chomanicová, 1988 [33]

Table 3 continued

Country	SWCT	Source
Spain	cover crops mulching buffer strips exclosure reduced tillage contour tillage soil amendment terraces	Gómez et al., 2009 [62]; Schnabel et al., 2001* [61] Albaladejo et al., 2000 [78]; Bautista et al., 1996 [74] Aspizua, 2003* [64]; Francia Martínez et al., 2006 [65]; Gómez et al., 2004 [63]; Martinez Raya et al., 2006* [65] Castillo et al., 1997 [78]; Martinez-Mena et al., 1999 [76] Bienes et al., 2006* [67-70]; Francia Martínez et al., 2006 [65]; Gómez et al., 2004 [78] Schnabel et al., 2001* [61] Ingelmo et al., 1998 [72]; Lopez-Bermudez et al., 1991 [77]; Romero-Díaz et al., 1999 [77] Llovet et al., 2009* [73]; Llovet, pers. comm. [73]; Martínez-Casasnovas and Ramos, 2006* [71]; Ramos and Martínez-Casasnovas, 2006, 2007 [71]; Sanchez et al., 1994 [75]; Williams et al., 1995* [66]
Sweden	cover crops no-tillage reduced tillage	Ulén and Kalisky, 2005 [7] Ulén and Kalisky, 2005 [7] Ulén and Kalisky, 2005 [7]
Syrian Arab Republic	cover crops exclosure reduced tillage terraces	Bruggeman et al., 2005 [89]; Masri et al., 2005 [89]; Shinjo et al., 2000 [90] Bruggeman et al., 2005 ^{TS} [89]; Masri et al., 2005 [89]; Ali, 2007* [88]
Tunisia	contour tillage	Kaabia, 1995 [83]
Turkey	strip cropping contour tillage	Köse and Taysun, 2002 [93]; Köse et al., 1996 [93]; Köse and Taysun, 2002^{TS} [93]; Köse et al., 1996^{TS} [93]; Oguz et al., 2006^{TS} [92,94-103]; Oguz, pers. comm. [92,94-103]
U.K.	mulching reduced tillage contour tillage geotextile	Brown, 1996 [16] Brown, 1996 [16]; Quinton and Catt, 2004 [17] Quinton and Catt, 2004 [17] Bhattacharyya et al., 2008, 2009 [15]; Mitchell et al., 2003 [15]

in plot data collection may also be linked to the substitution of field measurements studies by modelling studies. Nevertheless, interest in field-measured data on the effects of SWCTs remains high as shown by the increase of the number of publications in recent years. Recent publications also review and re-analyse plot data collected over long periods which were never published internationally (e.g. Rousseva et al., 2006).

Figure 2b shows that, while soil management (i.e. mainly tillage) and mechanical methods have been studied continually respecitively since 1956 (Jung and Brechtel, 1980) and 1962 (Pinczés, 1982), crop and vegetation management methods have only been studied significantly since 1984. This most likely reflects changing attitudes towards sustainable agricultural practices. The increase in interest in the effects of mechanical methods since 1990 is reflected in an increase in research on the effects of terracing on R_a and SL_a in reforested land (Llovet et al., 2009; Sanchez et al., 1994; Williams et al., 1995), and to the use of (biological) geotextiles (Bhattacharyya et al., 2008, 2009; Jankauskas et al., 2008; Kertesz et al., 2007; Mitchell et al., 2003).

Figure 3 shows the frequency distribution of the number of plots (#PL) and #PY over different plot slope lengths, slope gradients and annual precipitation depth. Soil management techniques (i.e. mostly tillage techniques) were mainly tested on gentle slopes while mechanical methods have been tested on steeper slopes. The frequency distribution of plot slope lengths and slope gradients is similar to the distribution for plots where no SWCTs are applied (Maetens et al., 2012), and is likely determined by experimental preferences, as the modal plot length and slope gradient are close to the standard (R)USLE plot which has a plot length of 22.13m and a slope gradient of 9% (Renard et al., 1997). The concentration of plot lengths and slope gradients around these values also implies that effects of plot length and slope gradient on SWCT effectiveness in reducing R_a and SL_a are difficult to assess for longer and steeper slopes.

Other important environmental data (e.g. soil characteristics such as texture, organic matter content) and climatological data (precipitation distribution and intensity) were not reported systematically in the reviewed literature and no global analysis of these variables was possible. Furthermore, #PL for some of the studied SWCTs is limited, so no further division is made according to land use types (e.g. cropland, tree crops, vineyards) as this would reduce the plot sample size. These data limitations imply that there is likely a large variability in R_a and

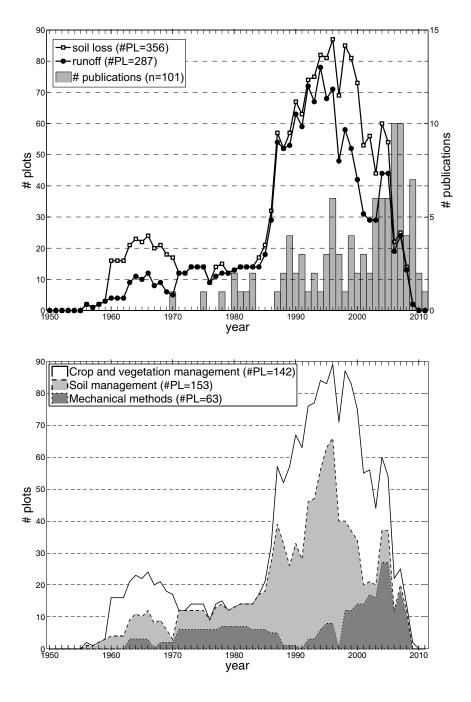


Figure 2: (a) Per-year evolution of the total number of erosion plots (# plots) set up to test soil and water conservation techniques (SWCTs) for which annual runoff and soil loss were recorded based on the individual plot (IP) database for Europe and the Mediterranean (Table 3). Grey bars represent the total number of publications (# publications) per publication year from which runoff and/or soil loss data were extracted. (b) Evolution of # plots per year for which annual soil loss and runoff were recorded on plots with a SWCT based on the IP database. The full black line indicates the total # plots, while dashed lines and grey tones indicate the relative proportions of the three major SWCT groups (Table 2). n= number of publications, #PL= number of plots.

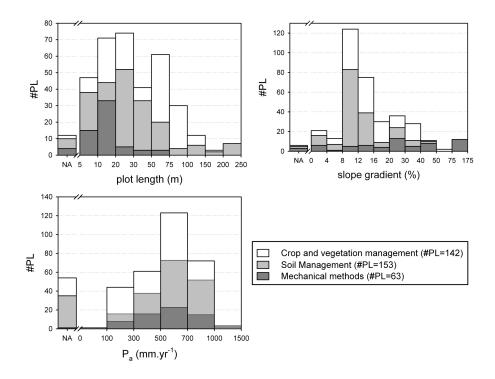


Figure 3: Frequency distribution of the number of plots (# PL) included in the individual plot database for Europe and the Mediterranean for which annual runoff and/or soil loss data are available as a function of: (a) plot length, (b) slope gradient and (c) annual precipitation depth (P_a). A distinction is made between the three major SWCT groups (Table 2). #PL: number of plots. NA: not available, i.e. plot length, slope gradient or Pa were not reported.

 SL_a which can not be explained with the environmental data currently available, and any observed trends should be interpreted with care, considering possible effects of these unknown factors.

4 Effectiveness of SWCTs in reducing annual runoff and soil loss

In order to obtain a general overview of the effects of SWCT, the frequency distributions of annual runoff coefficient (RC_a , %) and SL_a , from all plots where SWCTs were applied were compared to frequency distributions of RC_a and SL_a measured on plots where local conventional land management practices were applied (Figure 4). The latter were derived from a previously established erosion plot database (Maetens et al., 2012). This approach allows a broad assessment of the overall effects of SWCTs on RC_a and SL_a , but cannot be used to quantify the effectiveness of individual SWCTs. Therefore, all plots with SWCTs for which a paired plot under local conventional management existed (i.e. a plot on the same measuring site with the same characteristics but with application of the local conventional management instead of a SWCT) were selected from the full SWCT plot database (individual plot database, IP) and grouped together in a separate subset (paired plot database, PP).

4.1 Overall effectiveness of SWCTs

Fig. 4a does not show a clear effect of the application of SWCTs on the exceedance probability distribution of RC_a . While RC_a -values for plots with a (semi-)natural vegetation cover are clearly smaller than RC_a -values for bare plots or plots under conventional cropland practices,

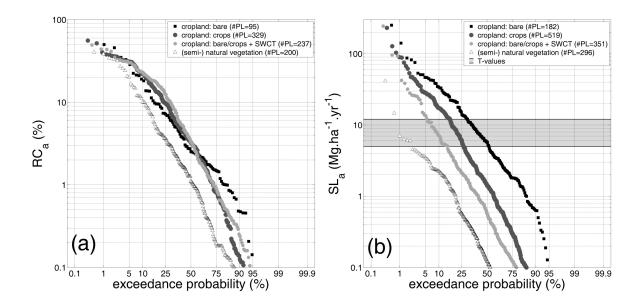


Figure 4: Exceedance probability distributions of (a) annual runoff coefficients (RC_a) and (b) annual soil loss (SL_a) for bare plots, cropland plots (annual crops, olive orchards and vineyards) without soil and water conservation techniques (SWCTs), plots on land with semi-natural vegetation (forest, shrubland and grassland; obtained from the plot runoff and soil loss database described in (Maetens et al., 2012), and all cropland plots with SWCT application (this study, Table 3). For SL_a , the range of tolerable soil losses (T-values, i.e. 5-12 $Mg \cdot ha^{-1} \cdot yr^{-1}$) as defined by (Montgomery, 2007b) is indicated in grey. #PL: number of plots.

this difference in RC_a is not observed for cropland plots where a SWCT was applied. On the other hand, Figure 4b shows that the presence of a crop on the land already substantially reduces the exceedance probability of any given SL_a rate compared to the exceedance probability for SL_a from bare plots, while the application of SWCTs reduces the exceedance probability even further.

The observed exceedance probabilities correspond to the trend observed by Montgomery (2007) in a worldwide study of SL_a rates. Nevertheless, the reduction of the exceedance probability of any given SL_a rate by the application of SWCTs is generally smaller in Europe and the Mediterranean (Figure 4b) than the reduction observed by Montgomery (2007). The latter report that the exceedance probability of a soil loss tolerance level (T-value) of 5 Mg·ha⁻¹·yr⁻¹ drops by 75% when SWCTs are applied, compared to SL_a rates on cropland without SWCTs (Table 3). However, for Europe and the Mediterranean (this study) a reduction of only 14% is observed (Table 3). The larger effect of SWCT application on SL_a rates in the global study (Montgomery, 2007b) is due to the fact that SL_a rates on cropland reported by Montgomery (2007b) are generally much larger than those observed in this study (Table 3). These larger SL_a rates may be attributed to the inclusion of more erosion-prone sites in the global study (Montgomery, 2007b) as compared to the erosion plot sites in Europe and the Mediterranean (this study). In addition, the reported SL_a rates from land where SWCTs were applied are in general somewhat smaller in the global study by Montgomery (2007b) than in this study (Table 3). This difference may be attributed to the comparatively smaller number of plots where SWCTs were applied in the study by Montgomery (2007b), and thus the lower probability of observing extreme SL_a rates in that study.

Table 3: Exceedance probabilities of annual soil loss on cropland without SWCTs and cropland with SWCTs for the worldwide study by Montgomery (2007b) and for Europe and the Mediterranean (this study), corresponding to two soil loss tolerance values (T). NA: not applicable(see also Figure 4).

	$T=5~\mathrm{Mg}$	$\cdot \text{ha}^{-1} \cdot \text{yr}^{-1}$	$T=12~\mathrm{Mg}$	$g \cdot ha^{-1} \cdot yr^{-1}$
	World	Europe and the Mediterranean	World	Europe and the Mediterranean
cropland without SWCT cropland with SWCT	$\begin{array}{c} \text{(Montgomery, 2007b)} \\ 85\% \\ 10\% \end{array}$	$\begin{array}{c} {}_{({\rm this\ study})} \\ 27\% \\ 13\% \end{array}$	$62\% \\ NA$	$\begin{array}{c} {\rm (this\ study)} \\ 16\% \\ 4\% \end{array}$
difference	75%	14%	NA	12%

4.2 Effectiveness per technique

To quantify the effectiveness of specific SWCTs, runoff ratios (RR) and soil loss ratios (SLR) were calculated for all paired plots (PP) in the database.

$$RR = \frac{R_{aSWCT}}{R_{aCONV}} \tag{Eq. 1}$$

$$SLR = \frac{SL_{aSWCT}}{SL_{aCONV}}$$
 (Eq. 2)

where:

 $R_{aSWCT} = \text{annual runoff (mm·yr}^{-1})$ measured on the plot with SWCT.

 $R_{aCONV} = \text{annual runoff (mm·yr}^{-1})$ measured on the plot with conventional practice.

 $SL_{aSWCT} = annual soil loss (Mg \cdot ha^{-1} \cdot yr^{-1})$ measured on the plot with SWCT.

 $SL_{aCONV} = \text{annual soil loss (Mg·ha}^{-1} \cdot \text{yr}^{-1})$ measured on the plot with conventional practice.

Except for deep tillage and terraces, the median effectiveness of all other SWCTs in reducing R_a is much smaller than for SL_a (Figure 5). Deep tillage is a technique aimed mainly at reducing runoff through increasing infiltration capacity (Chomanicová, 1988; Suchanic, 1987). For terraces, the higher median effectiveness in reducing R_a is explained by the fact that the terraces were established as water harvesting technique in an arid climate (Al-Seekh and Mohammad, 2009). The ranking of effectiveness for the individual SWCTs is different for R_a and SL_a reduction. Nevertheless, mechanical methods and crop and vegetation methods are generally more effective in reducing both R_a and SL_a than soil management methods. Exceptions to this trend are exclosures and strip cropping, which are less effective than the other SWCTs in the crop and vegetation management group. The relative order of SWCTs in terms of RR and SLR, underlines the large effectiveness of vegetation cover (cover crops, buffer strips), soil contact cover (mulching, geotextiles), or physical barriers (terraces, contour bunds) to reduce R_a and SL_a . While tillage methods are less effective in reducing R_a and SL_a than crop and vegetation management or mechanical methods, they may be better suited where competition for water is an issue (Unger and Vigil, 1998) or mechanical methods are not compatible with agricultural practices (e.g. when mechanized tillage would be obstructed by terraces or stone bunds). Comparing the total #PL for individual SWCTs with the median effectiveness of these techniques (Figure 6) also shows that larger efforts have been made for the assessment of the effectiveness of no-tillage, reduced tillage and contour tillage. However, these are generally not the most effective SWCTs (Figure 5). The limited #PL for the more effective SWCTs also limits the range of environmental conditions where these more promising SWCTs have been tested in Europe and the Mediterranean.

Weighted mean RR and SLR values were calculated for each SWCT (Table 4). In accordance with the central limit theorem (Maetens et al., 2012; Tijms, 2004), means were calculated by

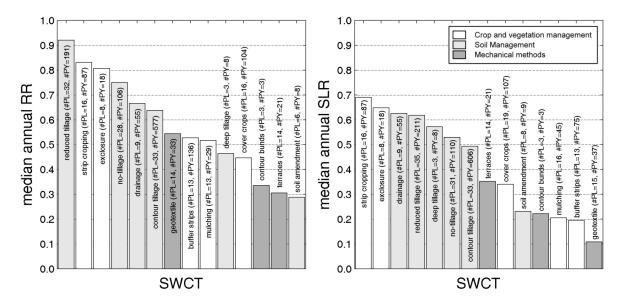


Figure 5: Median annual runoff ratios (RR, Eq. 1) and soil loss ratios (SLR, Eq. 2) for different soil and water conservation techniques (SWCTs) included in the paired plot database. SWCTs are ranked from left to right in order of increasing effectiveness. #PL: number of plots, #PY: number of plot-years.

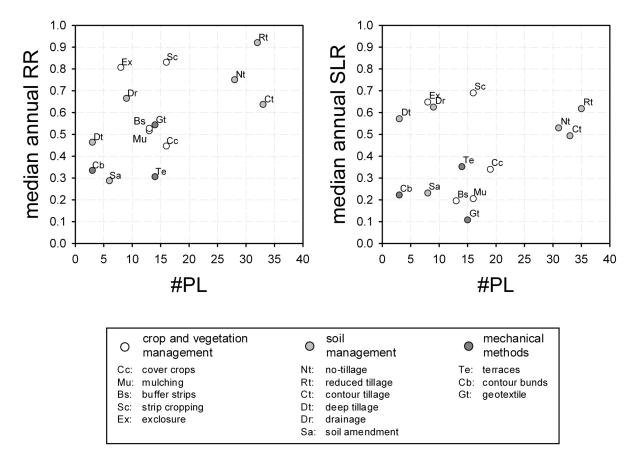


Figure 6: Median annual runoff ratio (RR, Eq. 1) and annual soil loss ratio (SLR, Eq. 2) for different SWCTs versus the number of plots (#PL) used to test the corresponding SWCT.

weighting the RR and SLR for each paired plot by the square root of the #PY. Figure 7 shows the weighted mean and distribution of SLR and RR for each SWCT. This figure illustrates that the range of SLR and RR for a given technique can be very large, indicating an important effect of local environmental and experimental conditions. Hence, the mean and median reduction ratios for a specific SWCT should be interpreted with caution as they may be subject to important variability. Furthermore, the frequency distributions of the reduction ratios are positively skewed, with a limited number of high RR and SLR values strongly affecting the weighted mean values (Figure 7). Nevertheless, both the median and weighted mean values show a similar pattern: most SWCTs are less effective in reducing runoff than in reducing soil loss.

Table 4: Number of plots (#PL), number of plot-years (#PY), weighted mean, median and coefficient of variation (CV) for annual runoff ratios (RR) and annual soil loss ratios (SLR) per soil and water conservation technique (SWCT) for all data in the paired plot database. Weighting for the mean is done according to square root of the number of plot-years. * Drainage: all reported results refer to surface R_a and SL_a only, hence not to runoff and soil loss measured from the pipe outflow.

			$\mathbf{R}\mathbf{R}$					\mathbf{SLR}		
SWCT	$\#\mathbf{PL}$	$\#\mathbf{PY}$	mean	mediai	n CV	$\#\mathbf{PL}$	$\#\mathbf{PY}$	mean	mediai	a CV
Crop and vegetation management										
cover crops	16	104	1.21	0.45	1.31	19	107	0.49	0.34	0.84
mulching	13	29	0.56	0.52	0.71	16	45	0.21	0.21	1.41
buffer strips	13	136	0.63	0.53	0.83	13	75	0.26	0.20	1.02
strip cropping	16	87	1.36	0.83	1.42	16	87	1.48	0.69	0.95
exclosure	8	18	0.83	0.81	0.43	8	18	0.54	0.65	0.68
Soil management										
no-tillage	28	106	0.87	0.75	0.78	31	110	0.72	0.53	1.00
reduced tillage	32	191	1.38	0.92	0.93	35	211	0.98	0.62	1.44
contour tillage	33	577	0.69	0.64	0.96	33	606	0.50	0.49	1.41
deep tillage	3	8	0.46	0.46	0.47	3	8	0.64	0.57	0.22
drainage*	9	55	0.61	0.67	0.18	9	55	0.63	0.63	0.28
soil amendment	6	8	0.38	0.29	0.75	8	9	0.36	0.23	0.95
Mechanical methods										
terraces	14	21	0.45	0.31	1.08	14	21	0.75	0.35	1.38
contour bunds	3	3	0.39	0.33	0.76	3	3	0.33	0.22	1.15
geotextile	14	33	0.71	0.54	0.84	15	37	0.16	0.11	1.62

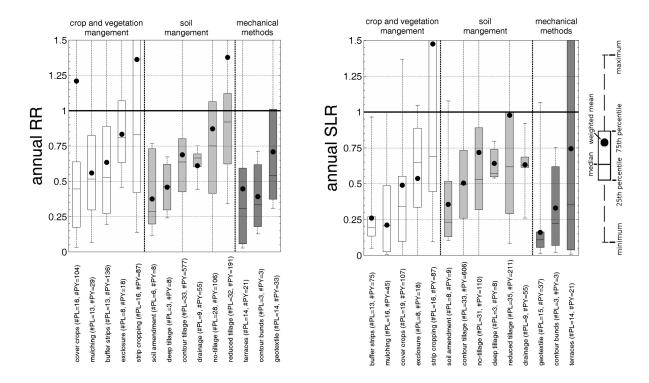


Figure 7: Frequency distribution and mean annual runoff ratio (RR, Eq. 1) and annual soil loss ratio (SLR, Eq. 2) for different soil and water conservation techniques (SWCTs) included in the paired plot database. SWCTs are ranked according to increasing median RR and SLR, within each of the 3 major SWCT groups (Table 2). #PL: number of plots, #PY: number of plot-years.

Relations between RR and SLR are given in Figure 8. Regarding crop en vegetation management techniques, mulching and buffer strips are generally more effective in reducing SL_a than cover crops, but the latter are slightly more effective in reducing R_a (Figure 8). With respect to tillage techniques, deep tillage seems less effective in reducing SL_a than contour tillage and no-tillage, but more effective in reducing R_a . For some plots where no-tillage or reduced tillage was applied, a reduction in SL_a but an increase in R_a was measured. This is probably due to increased soil sealing which reduces soil loss but limits infiltration and increases R_a . In these cases, soil conservation techniques are not necessarily soil and water conservation techniques. Hence, whether a SWCT reduces mainly R_a , SL_a or both is an important factor to consider when selecting the most suitable SWCT to address degradation problems on a specific site.

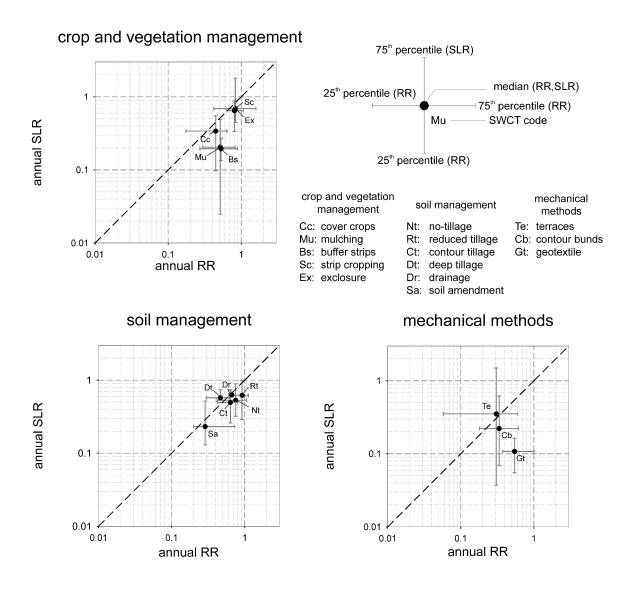


Figure 8: Relation between annual runoff ratio (RR, Eq. 1) and annual soil loss ratio (SLR, Eq. 2) for different soil and water conservation techniques (SWCTs) in the paired SWCT plot database. SWCTs plotting below the 1:1 line indicate soil loss reduction is stronger than runoff reduction.

5 Factors controlling the effectiveness of SWCTs

The large variability of RR and SLR values for individual SWCTs (Figure 7) prompts an analysis of the effect of environmental and experimental variables on the effectiveness of SWCTs. Due to a lack of data, this analysis was restricted to the variables reported in almost all publications from which data were extracted (Table 3): i.e. RC_a and SL_a from the reference plot, plot slope length, slope gradient and annual precipitation depth. R_a and SL_a are known to be strongly affected by soil characteristics (e.g. Bradford et al., 1987; Torri et al., 1997) as well as by rainfall distribution and intensity (e.g. Nearing et al., 2005). These factors may hence also affect SWCT effectiveness. However, these characteristics were insufficiently reported to allow a global analysis.

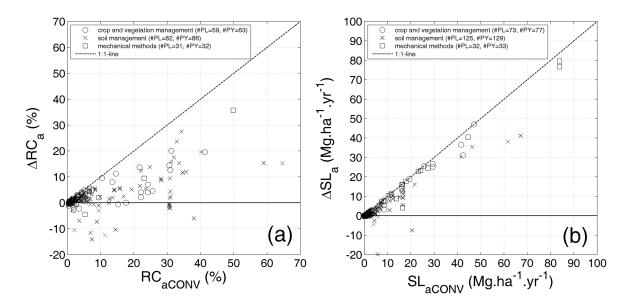


Figure 9: (a) Relation between the magnitude of annual runoff coefficients (RC_{aCONV}) measured on plots without soil and water conservation techniques (SWCTs) and the absolute reduction of annual runoff coefficients (ΔRC_a , Eq. 3) after applying a SWCT. (b) Relation between the magnitude of annual soil loss (SL_{aCONV}), measured on plots without SWCTs, and the absolute reduction of annual soil loss (ΔSL_a , Eq. 4) after applying a SWCT. #PL: number of plots, #PY: number of plot-years.

The effects of the RC_a and SL_a on the reference plot were investigated by calculating the absolute reduction of RC_a (Δ RC_a, %) and SL_a (Δ SL_a, Mg·ha⁻¹·yr⁻¹):

$$\Delta RC_a = RC_{aCONV} - RCa_{aSWCT} \tag{Eq. 3}$$

$$\Delta SL_a = SL_{aCONV} - SLa_{aSWCT} \tag{Eq. 4}$$

where:

 RC_{aSWCT} = annual runoff coefficient (%) measured on the plot with SWCT.

 RC_{aCONV} = annual runoff coefficient (%) measured on the plot with conventional practice.

Figure 9 shows that negative absolute reductions in SL_a and RC_a due to the application of SWCTs occur almost exclusively on plots where the RC_a and SL_a rates on the conventional plot are low. For RC_a , these observations generally occur at RC_{aCONV} rates below 20%. For SL_a , most negative ΔSL_a -values occur for SL_{aCONV} rates below 2 $Mg \cdot ha^{-1} \cdot yr^{-1}$. This is attributed to the large variability in SL_a and R_a intrinsic to plot studies (e.g. Nearing et al., 1999) when the absolute RC_a and SL_a rates are low. As a result, the estimated low effectiveness of SW-CTs for low absolute RC_a and SL_a values may be partly caused by the natural variability of of R_a and SL_a plot studies. Nevertheless, some negative ΔRC_a values were also observed for higher RC_a values (Figure 9). These values were derived from different studies where reduced tillage or no tillage techniques were applied (Chisci, 1989; Puustinen et al., 2005, 2007). Also for ΔSL_a , 8 negative values were noted for SL_{aCONV} rates larger than 2 $Mg \cdot ha^{-1} \cdot yr^{-1}$. Again, these points correspond to paired plot studies where tillage techniques (no-tillage, reduced tillage and contour tillage) were tested at different locations and under different land uses. This indicates that the lack of effectiveness is inherent to the tillage techniques, rather than depending on environmental conditions or measuring uncertainties.

To further explore relations between annual RR and SLR and plot slope length, slope gradient and annual precipitation depth for individual SWCTs, weighted regression equations of the type:

$$Y = aX^b (Eq. 5)$$

were calculated, whereby each observation in the regression was weighted by the square root of #PY corresponding to the plot measurement. The limited number of regressions for which the b-value is significantly different from 0 (α =0.05) indicates that the effect of plot slope length, plot slope gradient and Pa on SWCT effectiveness is limited (Table 5). For instance, no slope gradient and slope length effects on contour tillage effectiveness were found, even though both factors are used in the calculation of the P-factor for contour tillage (Renard et al., 1997). Nevertheless, all exponents (b) of the significant regressions are negative (Table 5), indicating that these SWCTs become more effective in more erosion-prone conditions (i.e. longer and steeper plot slopes and higher Pa). This is in line with results reported by Smets et al. (2008b,a) who observed that various soil surface cover types (i.e. organic and inorganic mulches, vegetation) become more effective in reducing SL on longer plots.

Table 5: Summary of the significant regressions (Y=aX^b, Eq. 5) between annual runoff ratio (RR), annual soil loss ratio (SLR) and plot slope length, plot slope gradient and annual precipitation for each of the soil and water conservation techniques (SWCTs). Only regressions significant at α =0.05 are given. #PL: number of plots, as regression coefficient, bs regression exponent, ps p-value; probability that the b-value is not significantly different from zero. r^2 : coefficient of determination.

variable	SWCT	#PL	a	b	p	\mathbf{r}^2			
		RF	Ł						
slope gradi	ient								
	mulching	13	3.28	-0.88	0.02	0.42			
	strip cropping	12	174.49	-1.83	0.01	0.55			
annual pre	ecipitation								
	contour tillage	29	43.57	-0.66	< 0.01	0.27			
	geotextile	14	$2.04 \text{x} 10^5$	-1.94	0.01	0.43			
		SLI	R						
plot length									
	cover crops	18	3.80	-1.06	0.04	0.25			
	terraces	14	527.62	-2.67	< 0.01	0.59			
annual pre	annual precipitation								
	mulching	14	861.36	-1.62	0.04	0.31			
	no-tillage	24	424.32	-1.02	0.04	0.18			
	terraces	14	$3.6 \text{x} 10^8$	-3.42	< 0.01	0.63			

The overall absence of clear relations between RR or SLR and plot slope length, plot slope gradient or Pa may be partly attributable to differences in other environmental and experimental conditions of the plot studies. These differences are hard to account for, because these conditions were rarely specified in the original data sources. For instance, rainfall distribution and intensity may have a large influence on RR and SLR, since R_a and SL_a often depend on a few low-frequency, high-magnitude events (e.g. de Figueiredo and Poesen, 1998). Unfortunately, insufficient rainfall distribution and intensity data were available to investigate potential relations with SWCT effectiveness. Likewise, soil characteristics may significantly control RR and SLR. However, no data were available to quantify this effect. Furthermore, the calculated regressions (Table 5) are strongly controlled by a few observations with a very high plot length or gradient. This further increases the uncertainty on the calculated relationships. As a result, our results remain inconclusive about the overall effects of environmental factors on the effectiveness of SW-CTs.

6 Temporal variability and trends in the effectiveness of SWCTs: the case of tillage techniques

Mean measuring period for plots in the PP database was 4.9 yrs. (st. dev.: 11.2 yrs, median: 2 yrs., mode: 1 yr.). Hence, most studies focus on the quantification of the effectiveness directly after application of the SWCT, while temporal variability and evolution of SWCT effectiveness over time is largely ignored.

To explore the effectiveness of SWCTs during consecutive years, 65 time-series of annual paired plot R_a and/or SL_a observations were selected from the PP database. At these plots, the effectiveness of no-tillage, reduced tillage or contour tillage under cropland was tested for a period of at least 3 consecutive years. To increase the number of available data, also studies from alsewhere where the effectiveness of no-tillage, contour tillage or reduced tillage under cropland was evaluated were included. These other studies include time-series from 18 paired plots from the United States (Bosch et al., 2005; Cullum et al., 2007, 2010; McCool et al., 2008; McDowell and McGregor, 1984; Schreiber and Cullum, 1998; Shipitalo and Edwards, 1998) and 12 plot data time-series from Australia (Freebairn and Wockner, 1986).

Figure 10 shows that there is considerable variability in RR and SLR for individual paired plots between the different years, independent of the time after SWCT application. The median trend of the RR and SLR over the consecutive years since the first application of the tillage techniques (Figure 10) shows that for no-tillage, RR are generally larger than SLR throughout the measuring period and tend to increase over the years following first application. Hence, the effectiveness of these techniques to reduce runoff decreases over time. No such trend is observed for SLR. This is most likely attributable to increasing surface sealing when the soil is not tilled for several years. While this is beneficial for SL_a reduction, it also reduces infiltration capacity of the soil and thus enhances R_a (Bradford et al., 1987). Similar, but less clear trends can be noted for reduced tillage and contour tillage (Figure 10). The number of sufficiently long time-series is limited however. As a result, these trends should be interpreted with caution as they may not be globally valid.

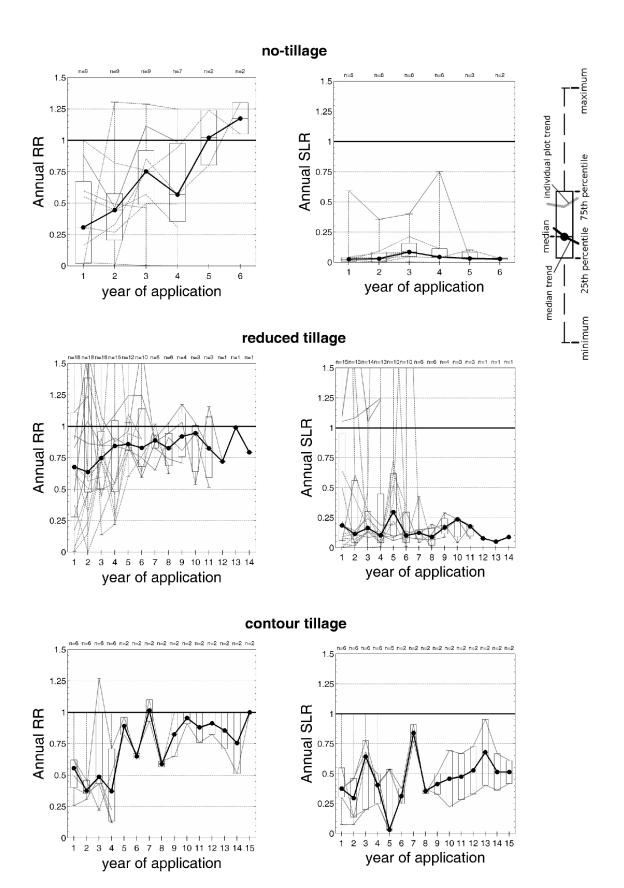


Figure 10: Frequency distribution (box plots) and median trend (black solid line) of annual runoff ratio (RR, Eq. 1) and annual soil loss ratio (SLR, Eq. 2) for consecutive years following the application of a soil and water conservation technique. All studies were conducted on plots under cropland. Thin grey solid lines indicate trends for individual plots in Europe and the Mediterranean, thin grey dashed lines indicate trends for individual plots in the U.S.A. or Australia. n = number of plots.

7 Conclusions

This study is the first to present a meta-analysis of field-measured plot data on the effectiveness of soil and water conservation techniques in reducing annual runoff (R_a) and annual soil loss (SL_a) for the Europe and the Mediterranean. Its results allow one to estimate the effectiveness of different SWCTs in reducing both R_a and/or SL_a . Crop and vegetation management (buffer strips, mulching, cover crops) and mechanical methods (geotextiles, terraces, contour bunds) are more effective in reducing R_a and SL_a than soil management techniques (no-tillage, reduced tillage, contour tillage). Nevertheless, large research efforts have been invested in these soil management techniques, while potentially more effective techniques remain relatively underresearched. A key finding of this study is that SWCTs are generally much more effective in reducing soil loss than in reducing runoff. This is important to consider, since reducing runoff and promoting infiltration may be a much bigger concern than reducing soil loss. Both for soil loss and runoff, the measured effectiveness for a given SWCT varies widely between different studies, indicating an effect of environmental factors. However, due to a lack of data the factors controlling this variability could not be clearly identified. Some of the studied soil and water conservation techniques become more effective in reducing R_a and SL_a rates in more erosion-prone conditions, i.e. when larger R_a and SL_a are measured on the plot with conventional practice, on longer and steeper slopes or in areas with a higher annual precipitation depth. However, observed relations between these factors and RR and SLR are weak and future studies should aim to better identify factors controlling the effectiveness of SWCTs in reducing R_a and SL_a , as this is a key requirement for their successful implementation.

Furthermore, the effectiveness of a SWCT may show important temporal variations. For no-tillage and, to a lesser extent, reduced tillage and contour tillage, clear indications were found that these techniques become less effective in reducing runoff after consecutive years of application. These longer-term trends, as well as the inter-annual variability of SWCT effectiveness are important factors to be further explored, as they will contribute to a better selection of these techniques to conserve both our soils and water resources. This will require studies evaluating the effectiveness of SWCTs over longer time periods (i.e. >3 years). Unfortunately, such studies are currently rare.

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