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1

2 Reversible flocculation of microalgae using magnesium hydroxide

3

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21

22 **Abstract**

23

24 Flocculation of microalgae is a promising low-cost strategy to harvest microalgae for bulk
25 biomass production. However, residual flocculants can interfere in further downstream
26 processes or influence biomass quality. In this study, a new concept is demonstrated based on
27 reversible magnesium hydroxide flocculation, using *Chlorella vulgaris* and
28 *Phaeodactylum tricornutum* as respectively a freshwater and a marine model species. We
29 show that flocculation was induced by precipitation of magnesium hydroxide at high pH (10
30 to 10.5). This resulted in a magnesium content of the microalgal biomass of 5% for *Chlorella*
31 and 18% for *Phaeodactylum*. After pre-concentration of the microalgal biomass by gravity
32 sedimentation, 95% of the precipitated magnesium hydroxide could be removed from the
33 biomass by mild acidification (pH 7 to 8). The pH fluctuations experienced by the microalgae
34 during flocculation/de-flocculation had no influence on biomass composition (FAME, total N
35 and P, carbohydrates, proteins, mineral content) and on the viability of microalgal cells.
36 Magnesium can thus be used as pH-dependent reversible flocculant for harvesting microalgae
37 in both marine and freshwater medium.

38

39

40 **Introduction**

41

42 Microalgae are a highly promising feedstock for production of biofuels. These unicellular
43 micro-organisms have a high areal productivity and produce biomass that is low in structural
44 compounds like cellulose or lignin and can thus be almost entirely valorised in a biorefinery
45 context, with the lipid fraction of the biomass being used for biodiesel production while the
46 remaining protein-rich residue can be used as animal feed [1]. Moreover, they do not compete
47 directly with food production by using agricultural crop area. Production of microalgae,
48 however, is a relatively energy-intensive process. To make microalgae competitive with
49 conventional agricultural crops, the cost and energy inputs of the production process need to
50 be reduced by at least an order of magnitude. One of the major challenges is situated in the
51 harvesting of the microalgae. Because microalgae cells are small (5-50 μm) and the biomass
52 concentration in the medium is low in case of open pond cultivation systems
53 ($< 1 \text{ g dry biomass L}^{-1}$), harvesting using centrifugation or membrane filtration is too costly
54 and energy-intensive[1, 2].

55

56 It is widely believed that flocculation has a lot of potential to reduce the cost and energy
57 demand of microalgae harvesting [3–6]. Using flocculation, microalgae can be pre-
58 concentrated 20-50 times using simple gravity sedimentation or flotation. The pre-
59 concentrated biomass can then subsequently be further dewatered using a mechanical method
60 such as filtration or centrifugation [7]. Because the bulk of the water has been removed during
61 the flocculation step, the cost and energy demand for further dewatering using mechanical
62 methods is reduced by more than an order of magnitude[3]. This has made microalgae
63 flocculation an active field of research in the past years. Many flocculation technologies that
64 have been widely used in other industries have been applied to microalgae harvesting. These

65 include metal coagulants, synthetic or natural polymer flocculants, electro-coagulation-
66 flocculation or lime softening [8, 9]. The main disadvantage of most flocculation
67 technologies, however, is that the harvested biomass becomes contaminated with the
68 flocculant, which may interfere with downstream processing. Borges et al., for instance,
69 showed that anionic polyacrylamides significantly influenced the fatty acid profile of lipid
70 extracts from marine microalgae [10]. Biomass contamination with a flocculant may also
71 interfere with valorisation of the protein-rich biomass residue remaining after extraction of
72 biofuels as animal feed.

73

74 When flocculation is used for harvesting microalgae it is part of a two-stage harvesting
75 process. During the first stage, flocculation is used in combination with gravity sedimentation
76 or flotation to pre-concentrate the biomass. During the second stage, the pre-concentrated
77 biomass is completely dewatered using a mechanical method like centrifugation. The ideal
78 flocculant for microalgae harvesting is a reversible flocculant that can be removed from the
79 microalgae biomass after the first stage, i.e. after pre-concentration of the biomass by
80 sedimentation or flotation. Magnetic iron oxide nanoparticles have been proposed for
81 reversible flocculation of microalgae. These nanoparticles are functionalized with a pH-
82 reversible surface charge. They interact with microalgae cells and cause flocculation at pH 12
83 but can be recovered multiple cycles from the pre-concentrated microalgae biomass at pH 2
84 [11]. The cost of commercially available magnetic nanoparticles is today high, it may become
85 a promising tool for harvesting microalgae when low-cost alternatives become available [12,
86 13]. Here, we propose an alternative method for harvesting microalgae using reversible
87 flocculation based on magnesium hydroxide.

88

89 One of the most promising low-cost methods for harvesting microalgae that has emerged in
90 the past years is pH-induced flocculation or the spontaneous flocculation of microalgae at
91 high pH. pH-induced flocculation is the result of precipitation of calcium or magnesium salts.
92 Calcium phosphate, for instance, precipitates at pH levels of 9 or above and can cause
93 effective flocculation of microalgae. The disadvantage of flocculation by calcium phosphate
94 precipitates is the high concentration of phosphate that is required. Indeed, more phosphate is
95 required for flocculating the biomass than for producing the microalgae biomass [14].
96 Magnesium hydroxide precipitates at a pH of 10-11 and can cause flocculation of microalgae
97 through charge neutralisation and/or by a sweeping mechanism. As most waters contain
98 sufficient quantities of magnesium, the only cost involved in magnesium hydroxide
99 flocculation is the cost of base required to increase the pH. This cost is low (about 18 \$ ton
100 dry biomass⁻¹ harvested) if a low-cost base such as lime is used [15]. Several recent studies
101 have highlighted the potential of pH-induced flocculation induced by magnesium hydroxide
102 precipitation as a low-cost harvesting method [3, 16, 17]. Nevertheless, this method of
103 harvesting results in a contamination of the biomass with magnesium hydroxide. Although
104 magnesium hydroxide is non-toxic, it is nevertheless desirable to remove it from the harvested
105 biomass.

106

107 Magnesium hydroxide precipitates above a pH of 10 to 11 (depending on the magnesium
108 concentration in the medium) and dissolves below this pH [18]. Magnesium hydroxide can
109 therefore theoretically be removed from the flocculated and pre-concentrated microalgae
110 biomass by means of acidification. In this way, magnesium hydroxide could be used as a
111 reversible flocculant: at high pH precipitation of magnesium hydroxide causes flocculation
112 while at low pH the magnesium hydroxide dissolves and can be removed from the biomass.
113 Whether this approach is applicable in practice depends on the pH required to dissolve the

114 magnesium hydroxide as well as the rate of dissolution during de-flocculation. If the pH
115 required is too low and/or the process takes too long, this approach is not feasible. Moreover,
116 the large pH fluctuations that are involved in the process may influence the quality in terms of
117 composition of the harvested microalgae biomass.

118

119 In this study, we evaluated the use of magnesium hydroxide as a reversible flocculant for
120 harvesting microalgae. We determined the optimal pH for flocculation and de-flocculation
121 and evaluated the impact of this reversible flocculation process on the microalgae biomass
122 composition and viability of the microalgae cells. Because magnesium concentrations differ
123 by almost two orders of magnitude between freshwater and seawater, the use of magnesium
124 hydroxide as a reversible flocculant was tested for a freshwater (*Chlorella vulgaris*) as well as
125 a marine (*Phaeodactylum tricornutum*) model microalgae species.

126

127

128 **Materials and Methods**

129

130 *Cultivation of C. vulgaris and P. tricornutum*

131

132 The freshwater green microalgae *Chlorella vulgaris* 211-11b (SAG) and the marine diatom

133 *Phaeodactylum tricornutum* 1055/1 (CCAP) were used as model species for studying

134 reversible magnesium flocculation. Both species have previously been used as a model

135 species in pH-induced flocculation studies [15, 19].

136 *C. vulgaris* and *P. tricornutum* were both cultured in batch mode in 10 L plexiglass bubble

137 column photobioreactors (20 cm diameter) using modified Wright's Cryptophyte medium

138 (Table 1) [20]. Additional magnesium and calcium were added to the cultivation medium of

139 both species in concentrations that are typically found in surface and in marine waters. The

140 reactors were aerated with 0.2 μm filtered air (5 L min^{-1}) and pH was maintained at 8.5

141 through pH-controlled addition of carbon dioxide to the air flow. The culture was irradiated

142 from two sides with daylight fluorescent tubes (Osram GroLux Sylvania, Germany), giving a

143 photon flux of $60 \mu\text{mol m}^{-2} \text{ s}^{-1}$ at the surface of the reactor. Algal biomass was monitored by

144 measuring absorbance at 750 nm. Absorbance measurements were calibrated against dry

145 weight, which was in its turn determined gravimetrically on pre-weighed GF/F glass fiber

146 filters [21]. Flocculation experiments were carried out at the end of the exponential growth

147 phase at day 6 of cultivation.

148

149 *Flocculation experiments*

150

151 We first determined the degree of contamination of the microalgae biomass by magnesium at

152 different pH levels during pH-induced flocculation by magnesium hydroxide. Flocculation

153 experiments were carried out in jars containing 100 mL of microalgae broth and mixing was
154 achieved by magnetic stirring. pH was adjusted by addition of 0.5 M NaOH in concentrations
155 ranging from 0 to 4mM . NaOH was dosed during 10 min of intensive mixing at 1000 rpm.
156 Then, the suspensions were mixed gently (250 rpm) for another 20 min while pH of each
157 treatment was monitored, followed by 30 min of settling. The flocculation efficiency was
158 calculated based on changes in the optical density (measured at 750 nm) prior to pH
159 adjustment (OD_i) and after settling (OD_f). The flocculation efficiency, or the percentage of
160 microalgae biomass removed from suspension, was calculated as:

$$161 \text{ Microalgae flocculation efficiency} = \frac{OD_i - OD_f}{OD_i}$$

162 To quantify the amount of magnesium that had precipitated during pH-induced flocculation,
163 we compared concentrations of magnesium in the supernatant and in the biomass pellet after
164 flocculation. Subsamples from supernatant and biomass pellet were centrifuged at 3,005 g and
165 stored at -20°C.

166

167 Samples from the supernatant were diluted and acidified to obtain a final concentration of
168 0.1 M HNO_3 (addition of 0.1 mL 5M HNO_3 to a 5 mL sample). The wet samples from the
169 biomass pellet were acidified with 10ml concentrated HNO_3 (70%) and after 48 hours the
170 clear solution was diluted 20 times with deionized water. Magnesium concentrations were
171 measured using ICP-MS (Agilent 7700x ICP-MS).

172

173 *De-flocculation experiments*

174

175 To evaluate the amount of magnesium that could be redissolved after acidification, triplicate
176 1 L suspensions of microalgae were flocculated. Based on previous results, the optimal
177 dosage of base in order to maximize flocculation efficiency and minimize the magnesium

178 content in the biomass was 4 mM of NaOH for *Chlorella* and 5 mM of NaOH for
179 *Phaeodactylum*. After 20 min of stirring at 250 rpm, the suspension was allowed to settle 30
180 min in 1 L Imhoff cones and the settled biomass pellet and the supernatant were separated by
181 decantation. 50 mL of the biomass pellet was subsequently de-flocculated by acidification
182 with 5 M HCl until pH 6, 7 and 8 and continuously stirred at 250 rpm for 30 and 60 minutes.
183 Subsamples were taken and prepared as previously described to compare magnesium
184 concentrations in supernatant and biomass pellet before, after flocculation and after de-
185 flocculation.

186

187 *Impact of flocculation and de-flocculation on the composition of the harvested biomass*

188

189 Flocculation (same conditions as above) and de-flocculation (pH 8 for 30 min) was carried
190 out on triplicate 1 L suspensions in order to evaluate the influence of the process on the
191 composition of the harvested biomass. After flocculation and de-flocculation, the biomass
192 pellet was separated from the supernatant by centrifugation, freeze-dried and stored at -80°C
193 until further analysis. Because the increase in pH to induce flocculation may not only result in
194 precipitation of magnesium hydroxide but also of other minerals (e.g. calcium phosphates,
195 calcium carbonate or calcium ammonium phosphate), we determined the content of
196 magnesium, calcium and phosphorus using ICP-MS as previously described. Carbohydrates,
197 fatty acid methyl ester (FAME) and protein content were compared with biomass before
198 flocculation as a control treatment to evaluate the influence of flocculation de-flocculation on
199 the biochemical composition of the biomass. The carbohydrates were extracted and measured
200 using the phenol-sulphuric acid method [22, 23]. FAME content was analysed using a direct
201 transesterification method [24] and the obtained FAMES were separated by gas
202 chromatography (GC) with cold on-column injection and flame ionisation detection (FID)

203 (Thermo Scientific Trace GC Ultra) using a Grace EC Wax column (length: 30 m, ID
204 0.32 mm, film: 0.25 μm). The used time–temperature program was: 70–180 $^{\circ}\text{C}$ (5 $^{\circ}\text{C}/\text{min}$),
205 180–235 $^{\circ}\text{C}$ (2 $^{\circ}\text{C}/\text{min}$), 235 $^{\circ}\text{C}$ (9.5 min). Fatty acid identification was performed using
206 standards containing a total of 35 different FAMES (Nu-check). Peak areas were quantified
207 with Chromcard for Windows software (Interscience). For analysis of the nitrogen content,
208 the biomass was digested using an alkaline persulphate digestion [25]. The nitrogen content in
209 the digestate was measured as nitrate and using Hach-Lange standard kits LCK 339 (Merck,
210 Darmstadt, Germany). The protein content of the biomass was calculated by multiplying the
211 nitrogen content (% on dry weight) with the general 6.25 conversion factor [26]. Results were
212 statistically analysed using a one-sample t-test with a level of significance of 0.05 (Sigmaplot
213 Systat Software).

214

215 *Cell viability assessments*

216

217 The effect of pH shifts on cell viability was first studied in a cultivation experiment using
218 modified Wright's Cryptophyte medium (Table 1). Deflocculated biomass (60 min at pH 8)
219 was compared to a control inoculum both for *Chlorella* and *Phaeodactylum*. Biomass density
220 was monitored daily spectrophotometrically (750nm) for 7 days.

221

222 Secondly, the maximum quantum yield of photosystem II (ratio of variable versus maximal
223 fluorescence; Fv:Fm) was measured to evaluate the influence of harvesting by
224 flocculation/de-flocculation on the photosynthetic apparatus using an AquaPEN PAM
225 fluorometer (Photon Systems Instruments). This parameter is a sensitive indicator of stress
226 experienced by microalgae and is often used for evaluating toxicity of substances towards
227 microalgae [27]. Quantum yield (n=3) was compared for *Chlorella* and *Phaeodactylum* before

228 flocculation, after 30 min of sedimentation and after de-flocculation (60 min at pH 8).
229 Measurements were conducted after 20 min of dark adaptation of the microalgae. Cells
230 treated with 15% H₂O₂ for 30 min were used as negative control. Results were statistically
231 analysed using a two way ANOVA and a pairwise multiple comparison (Tukey Test) with a
232 level of significance of 0.05 (Sigmaplot Systat Software).

233

234 Finally, cells were stained using Evan's Blue for *Chlorella* and *Phaeodactylum* before
235 flocculation, after 30 min of sedimentation and after de-flocculation (60 min at pH 8). This
236 was compared to a positive (no treatment) and a negative control (15% H₂O₂ for 30 min).
237 Cells were examined using light microscopy (Olympus BX 51).

238

239 **Results and discussion**

240

241 *Magnesium precipitation during flocculation*

242

243 We first experimentally quantified the degree of contamination of the biomass by magnesium
244 during pH-induced flocculation of *Chlorella* and *Phaeodactylum* (Fig 1). The biomass
245 concentration was about 0.1 g L⁻¹ in the *Chlorella* culture and 0.5 g L⁻¹ in the *Phaeodactylum*
246 culture.

247

248 In both species, flocculation occurred after addition of about 1 mM sodium hydroxide. At the
249 onset of flocculation, the pH was 10.7 for *Chlorella* and 10.3 for *Phaeodactylum*. In the
250 *Chlorella* experiment, about 0.25 mM of the initial 1 mM magnesium had precipitated from
251 solution at the onset of flocculation. Continued addition of sodium hydroxide did not change
252 the flocculation efficiency but resulted in an increased precipitation of magnesium until all
253 magnesium was removed from solution (Fig 1c). In the *Phaeodactylum* experiment, about 1.2
254 mM of the initial 50.15 mM magnesium had precipitated from solution at the onset of
255 flocculation (result not shown). Increasing the dose of sodium hydroxide resulted in increased
256 precipitation of magnesium, up to 2.5 mM magnesium at a dosage of 4 mM sodium hydroxide
257 (Fig 1f). This massive precipitation of magnesium resulted in the formation of a large amount
258 of white sludge. The maximum flocculation efficiency was 90% in *Chlorella* but only 73% in
259 *Phaeodactylum*. The flocculation efficiency in *Phaeodactylum* was probably underestimated
260 due to the formation of a fine white precipitate during flocculation, most likely calcium
261 carbonate. This resulted in a high residual absorbance of the supernatant despite the fact that
262 most cells had clearly flocculated.

263

264 We can assume that magnesium precipitated as magnesium hydroxide at high pH [15, 28].
265 Magnesium hydroxide is positively charged (point of zero charge 11.5) and can cause
266 flocculation through charge neutralisation and/or by a sweeping flocculation mechanism. The
267 flocculation mechanism is in that perspective similar to flocculation caused by aluminium
268 sulphate or ferric chloride, which also form positively charged hydroxides when dissolved in
269 water [29]. At the onset of flocculation, the concentration of magnesium in the microalgae
270 biomass was 4.7 % for *Chlorella* and 5.8 % for *Phaeodactylum* (Fig 1c and 1f), but this
271 increased rapidly as when the dosage of sodium hydroxide increased. This magnesium
272 hydroxide would end up in the harvested biomass and would result in a very high total
273 mineral content of the harvested biomass. In the current commercial production of
274 microalgae, guidelines state that the total mineral content should remain below 10 % [30].
275 With a magnesium content of approximately 5 %, the magnesium hydroxide content alone
276 will already result in a mineral content higher than 10 %.

277

278 *Dissolution of precipitated magnesium after de-flocculation*

279

280

281 We subsequently evaluated whether the precipitated magnesium hydroxide could be removed
282 from the flocculated biomass by means of de-flocculation by mild acidification. In this
283 experiment, the biomass concentration was 0.30 g L⁻¹ for *Chlorella* and 0.35 g L⁻¹ for
284 *Phaeodactylum*.

285

286 Prior to pH-induced flocculation, a magnesium concentration of only 0.6 % for *Chlorella* and
287 1.26 % for *Phaeodactylum* was found in the biomass pellet (Fig 2, BF). Magnesium in
288 microalgae is mostly associated with chlorophylls, with ATP and with enzymes involved in
289 phosphate metabolisms (ATP, nucleic acids) and typically amounts to a magnesium content of

290 about 0.5 %, which is close to what we measured in *Chlorella* [31, 32]. The higher
291 magnesium content of *Phaeodactylum* may be due to the fact that some magnesium hydroxide
292 had precipitated during cultivation as a result of the high pH of the cultures. Due to the higher
293 magnesium concentration in seawater when compared to freshwater, magnesium precipitation
294 is more likely to occur in seawater than in freshwater medium.

295

296 After addition of sodium hydroxide to induce flocculation, the amount of magnesium in the
297 biomass pellet increased substantially to 5 % in *Chlorella* and 18 % in *Phaeodactylum* (Fig 2,
298 AF). Magnesium concentration in the medium decreased from 1 mM before to 0.48 mM after
299 flocculation in the *Chlorella* experiment, indicating that almost half of the magnesium had
300 precipitated from solution. Because of the extremely high concentration of magnesium in the
301 seawater medium, pH-induced flocculation had no measurable effect on the magnesium
302 concentration in the medium in the *Phaeodactylum* experiment.

303

304 The flocculated microalgae were separated from the culture medium and acidified to re-
305 dissolve the magnesium that had precipitated during pH-induced flocculation. Three pH levels
306 were tested (6, 7 and 8) and the quantity of magnesium in the solution as well as in the
307 biomass pellet was measured after 30 and 60 minutes (Fig. 2). For the pH 6 and 7 treatments
308 in *Chlorella*, the magnesium in the biomass pellet after de-flocculation was the same as before
309 pH-induced flocculation, indicating that all magnesium that had precipitated was effectively
310 re-dissolved. For the pH 8 treatment, the magnesium content of the biomass pellet was
311 slightly higher than the initial magnesium content (0.8% instead of 0.6%). Nevertheless, this
312 indicates that 95% of the magnesium that had precipitated during pH-induced flocculation
313 could be re-dissolved by de-flocculation to pH 8. The de-flocculation time (30 or 60 minutes)
314 had no effect on the magnesium content in the biomass in the *Chlorella* experiment. For

315 *Phaeodactylum*, the magnesium content of the biomass pellet was reduced to 0.9% at pH 6,
316 1.1% at pH 7 and 1.5% at pH 8. The magnesium content was slightly lower after 60 than after
317 30 minutes of de-flocculation. The fact that the magnesium concentration in the biomass
318 pellet was in some cases lower than before pH-induced flocculation may be due to the fact
319 that acidification removed magnesium that had already precipitated in the *Phaeodactylum*
320 culture prior to flocculation.

321

322 These results indicate that the magnesium that precipitates during flocculation at high pH can
323 easily be re-dissolved and removed from the harvested biomass. Even a mild de-flocculation
324 by acidification at pH 8 for 30 minutes is sufficient to remove 95% of the magnesium from
325 the flocculated and pre-concentrated biomass. Magnesium hydroxide is a mineral that is easily
326 dissolved. Moreover, the ionic composition of the medium or the presence of organic
327 compounds in the medium has little effect on the dissolution kinetics of magnesium
328 hydroxide[33]. To acidify the medium to pH 8, only 0.75 mmol of hydrochloric acid was
329 added for *Chlorella* and 2.35 mmol for *Phaeodactylum*.

330

331 *Influence of flocculation – de-flocculation on biomass quality*

332

333

334 The rapid pH shifts experienced by microalgae during pH-induced flocculation and de-
335 flocculation may influence the viability of the microalgae cells as well as the biochemical
336 composition of the microalgae biomass.

337

338 Although it is not essential that microalgal cells remain viable during harvesting, viability
339 indicates that cell integrity remains intact and this in turn implies that the cell content is not
340 released into the medium during harvesting. In our experiments, cell viability did not appear

341 to be affected by the pH shifts. When microalgae cells were flocculated at pH 10.9 and
342 magnesium hydroxide was re-dissolved at pH 8 for 60 min., the cells immediately resumed
343 growth when transferred to fresh culture medium (Fig 3). For both species, growth rates were
344 comparable between the flocculated cells and the control cells. The quantum yield of
345 photosystem II of the cells or Fv:Fm, which is a sensitive indicator of stress in plants and
346 microalgae [27], did not significantly differ before flocculation, immediately after
347 flocculation and after re-dissolution of magnesium ($P > 0.270$), whereas a positive control
348 consisting of hydrogen peroxide treatment resulted in a large and significant decrease in
349 Fv:Fm ($P < 0.001$) (Fig 4). Finally, Evan's Blue staining showed that the cell membrane
350 remained intact throughout the flocculation – de-flocculation treatment in both *Chlorella* and
351 *Phaeodactylum* (Fig 5), while H₂O₂ treated cells again showed a clear blue staining of the
352 cytoplasm indicating a decrease of cell membrane integrity (Fig 5 (B,E)).

353

354 The carbohydrate, FAME, protein content of the biomass was also compared before
355 flocculation and after a flocculation – de-flocculation treatment (Table 2). In general, no
356 significant differences were observed. The pH increase required to induce precipitation of
357 magnesium hydroxide may also result in precipitation of other minerals from solution, such as
358 calcium phosphate and calcium carbonate[18]. These minerals may dissolve less easily during
359 de-flocculation by acidification and may thus accumulate in the harvested biomass. Calcium,
360 magnesium and phosphorus concentrations in the biomass were compared before and after the
361 flocculation – de-flocculation treatment (Table 2). We did not observe an significant increase
362 in mineral concentration of any of these minerals. We did, however, observe a small but
363 significant decrease in the calcium content for *Chlorella* (Table 2, $P = 0.008$). Possibly, some
364 precipitation of calcium carbonate had occurred in the cultures as a result of photosynthetic

365 increase in pH and this mineral was dissolved during de-flocculation of the pre-concentrated
366 biomass.

367

368 *Magnesium hydroxide as a reversible flocculant for freshwater as opposed to marine*
369 *microalgae*

370

371 Our results show that magnesium hydroxide can be used as a reversible flocculant for both
372 freshwater and marine microalgae. Addition of base causes precipitation of magnesium
373 hydroxide and this acts as an effective flocculant. Previous work has shown that the biomass
374 can be pre-concentrated about 50 times using magnesium hydroxide flocculation [28]. After
375 flocculation and pre-concentration of the biomass by gravity sedimentation, the precipitated
376 magnesium hydroxide can be removed from the pre-concentrated biomass by means of de-
377 flocculation by mild acidification. For both species, a decrease in pH to 8 for 60 min was
378 sufficient to remove 95% of the precipitated magnesium from the biomass.

379

380 Freshwaters have a relatively low concentration of magnesium, ranging from 0.1 mM in soft
381 water to 1 mM in hard water. Because the magnesium content of freshwater is low, the pH has
382 to be raised to a relatively high level before flocculation can be induced (pH 10.6 or above).

383 After pH-induced flocculation, the magnesium content in the biomass was around 5 %.

384 Overdosing of base during flocculation results in a slight increase of magnesium content in
385 the biomass, but never excessively because of the low magnesium content of the water.

386 Because the magnesium concentration in freshwaters is low, a large proportion of the
387 magnesium disappears from solution during flocculation, about 0.25 – 0.5 mM magnesium.

388 This depletion of magnesium from solution may pose problems when the culture medium is
389 recycled, as the magnesium concentration in the recycled medium may be too low to induce

390 flocculation. Recycling of the culture medium is essential to minimize the need for water
391 during production of microalgae biomass [34]. It is therefore important to recycle the
392 magnesium that is consumed during flocculation. When magnesium is dissolved during the
393 de-flocculation, it can be separated from the pre-concentrated biomass and returned to the
394 culture medium to be used in a second round of flocculation. In our experiment, more than
395 95 % of the magnesium could be recycled.

396

397 In seawater, magnesium is the second most abundant cation after sodium. Because the
398 magnesium concentration is very high in seawater, flocculation starts at a lower pH when
399 compared to freshwater medium [17, 28]. Some precipitation of magnesium hydroxide and
400 even some flocculation may already occur spontaneously due to photosynthetic rise of pH in
401 the culture [35]. Overdosing of base can cause massive precipitation of magnesium and
402 results in a large sludge volume and a very high magnesium content of the harvested biomass
403 and this should be avoided [28]. The magnesium hydroxide that has precipitated can be easily
404 removed from the biomass by mild acidification during the de-flocculation stage. The de-
405 flocculation step has an additional advantage in that the calcium content of the harvested
406 biomass is reduced. Precipitation of magnesium during flocculation barely lowers the
407 magnesium concentration in the medium. Therefore, it is not essential to return the
408 magnesium that is dissolved during the de-flocculation step to the culture.

409

410 In this study, flocculation was induced by addition of sodium hydroxide and de-flocculation
411 was demonstrated using hydrochloric acid. In the most optimal scenario, flocculation of
412 microalgae (0.5 g L^{-1}) was achieved after addition of 1.5 mM sodium hydroxide. This
413 corresponds with a dose of 0.12 ton sodium hydroxide ton^{-1} microalgal biomass, resulting in a
414 cost of 42 \$ ton^{-1} biomass (Table 3). However, this cost can be significantly reduced to 17 \$

415 ton^{-1} biomass if calcium hydroxide (slaked lime) is used instead of sodium hydroxide [15].
416 Based on our results, only 0.05 ton hydrochloric acid ton^{-1} biomass is needed for de-
417 flocculation. However, this would increase the cost of harvesting by at least 40 \$ ton^{-1}
418 biomass. The use of nitric acid could be an interesting alternative since the nitrate in nitric
419 acid could be used as nitrogen source for the microalgae when the medium is recycled. The
420 dosage of nitric acid that would be required for de-flocculation (0.75- 2.5 mM) corresponds
421 with the recommended nitrogen content of the microalgae cultivation medium (Table 1).

422 **Conclusion**

423
424 Our results show that magnesium can be effectively used as a pH-dependent reversible
425 flocculant for harvesting microalgae. At high pH, precipitation of magnesium hydroxide
426 results in flocculation of microalgae. At low pH, the magnesium is dissolved and removed
427 from the microalgae biomass. Thus, contamination of the harvested microalgae biomass by
428 the flocculant is avoided. The pH-shift required for flocculation and de-flocculation does not
429 affect the viability of the microalgae, nor the biochemical composition of the biomass. In
430 freshwater medium, the de-flocculation stage allows recycling of the magnesium and thus
431 avoids depletion of magnesium from the culture medium.

432

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434

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439 this work.

440

441 **Figures**

442

443 **Fig 1: Flocculation efficiency, pH after sodium hydroxide dosage and magnesium**
444 **concentration in supernatant (S) and biomass pellet (P) for *Chlorella vulgaris* (A,B,C)**
445 **and *Phaeodactylum tricornutum* (D,E,F)**

446

447 **Fig 2: Magnesium concentration in supernatant and biomass pellet before flocculation**
448 **(BF), after pH-induced flocculation (AF) and after 30 and 60 min of de-flocculation at**
449 **pH 6, 7 and 8 for (A) *Chlorella vulgaris* and (B) *Phaeodactylum tricornutum*.**

450

451 **Fig 3 : Comparison of growth using flocculated – de-flocculated biomass compared to a**
452 **control inoculum for (A) *Chlorella vulgaris* and (B) *Phaeodactylum tricornutum***

453

454 **Fig 4: Quantum Yield of PSII for *Chlorella vulgaris* and *Phaeodactylum tricornutum***
455 **(*Ptri*) before flocculation (BF), after flocculation (AF), after de-flocculation (60 min at**
456 **pH 8) and a negative control of microalgae with 15% H₂O₂**

457

458 **Fig 5: Evan's Blue staining essay: *Chlorella vulgaris* (A) no treatment, (B) H₂O₂ for 30**
459 **min and (C) after de-flocculation 60 min pH 8 and *Phaeodactylum tricornutum*: (D) no**
460 **treatment, (E) H₂O₂ for 30 min and (F) after de-flocculation 60 min pH 8**

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462 **References**

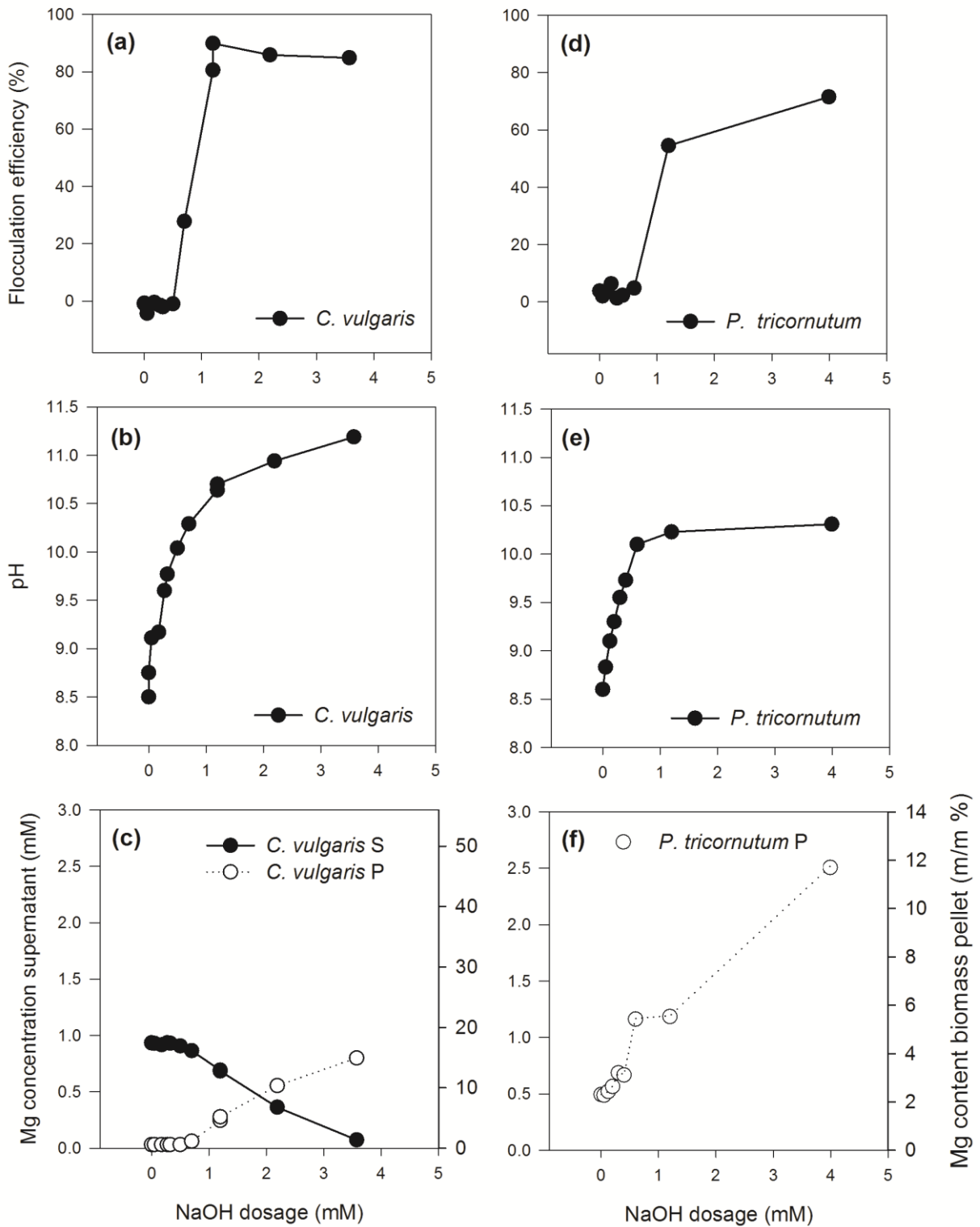
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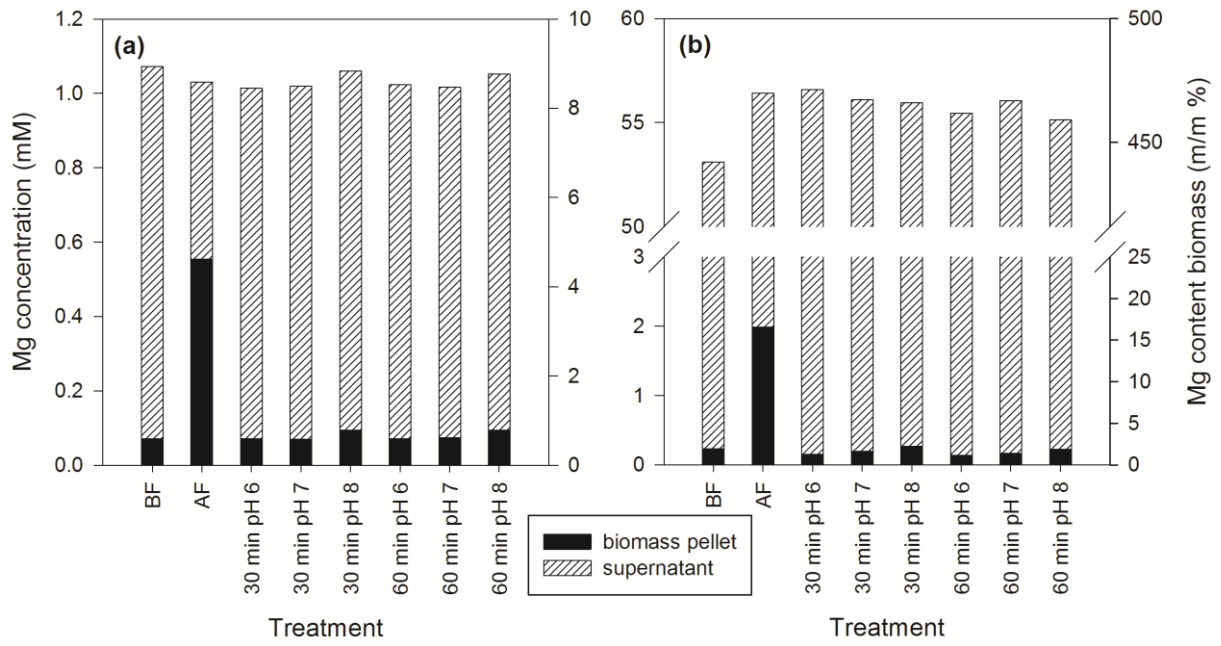
560 **Figure 1**



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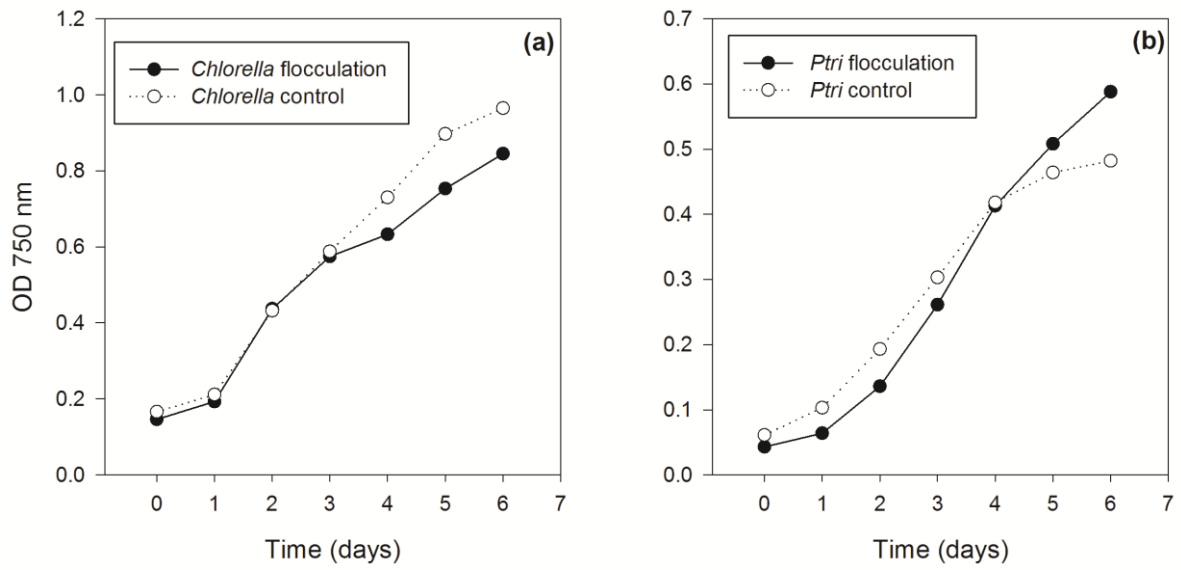
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563 **Figure 2**



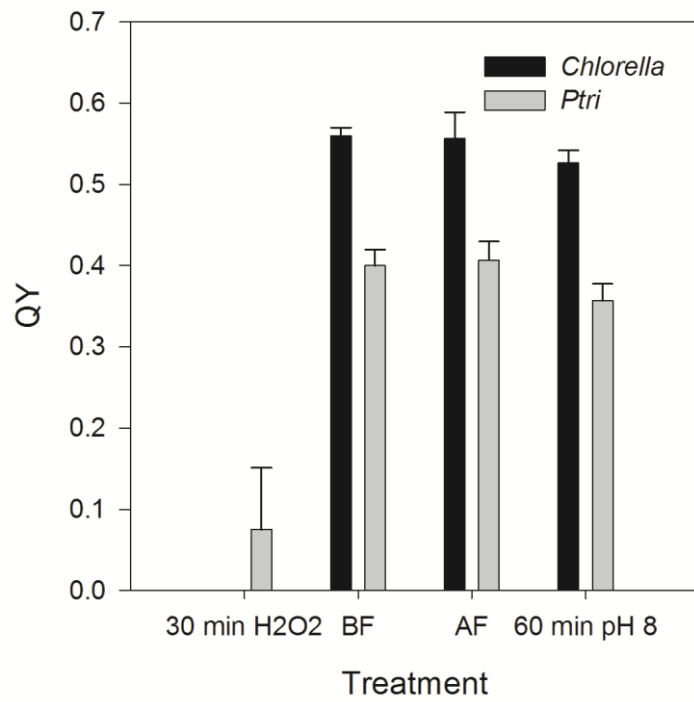
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565 **Figure 3**



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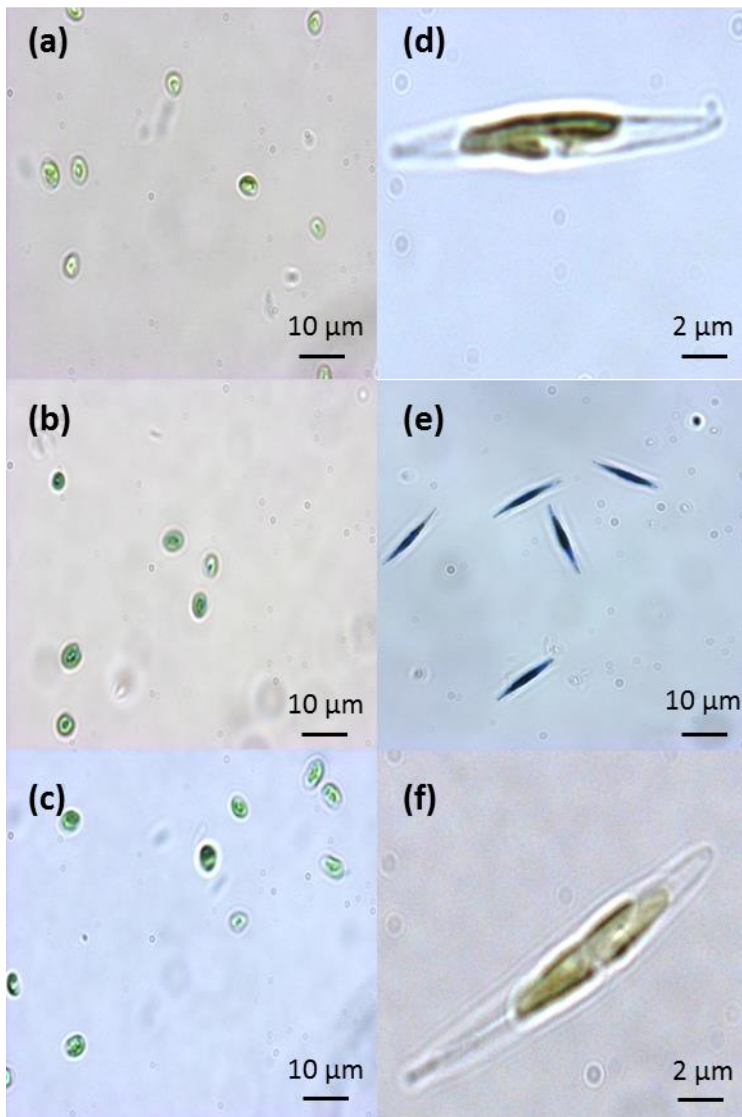
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568 **Figure 4**

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570

571 **Figure 5**



572

573

574 **Tables**

575

576 **Table 1. Concentrations of the main ions in the cultivation medium for *Chlorella vulgaris***
577 **(freshwater) and *Phaeodactylum tricornutum* (seawater) in batch cultivation mode**
578

	<i>C. vulgaris</i> (mM)	<i>P. Tricornutum</i> (mM)
Cl ⁻	1.7	633.3
Na ⁺	1.9	569.3
Mg ²⁺	1.0	50.15
Ca ²⁺	2.7	10
K ⁺	0.3	0.3
NO ₃ ⁻	1	1
PO ₄ ³⁻	0.05	0.05
SO ₄ ²⁻	1.3	25

579

580

581 **Table 2: Comparison of biomass composition in terms of carbohydrates, total FAME,**
582 **proteins, Ca, Mg and P (m/m % dry weight) before (C) and after flocculation – de-**
583 **flocculation (FDF) for *Chlorella vulgaris* and *Phaeodactylum tricornutum***
584

content (%)	<i>Chlorella vulgaris</i>			<i>Phaeodactylum tricornutum</i>		
	C	FDF	P-value	C	FDF	P-value
Carbohydrates	31.2	27.9 ± 1.85	0.489	13.9	14.5 ± 1.09	0.445
FAME	24.3	25.8 ± 1.16	0.155	12.1	13.6 ± 0.69	0.066
Proteins	22.4	22.9 ± 0.57	0.284	27.8	28.8 ± 1.75	0.404
Ca	0.22	0.11 ± 0.02	0.008	0.55	0.32 ± 0.13	0.093
Mg	0.42	0.57 ± 0.12	0.162	0.87	0.77 ± 0.17	0.407
P	0.57	0.51 ± 0.08	0.324	0.24	0.24 ± 0.01	0.240

585

586

587 **Table 3: Cost calculation based on optimal scenario for reversible magnesium**
588 **flocculation**
589

Flocculation	
Biomass concentration (g L ⁻¹)	0.5
Dose NaOH (ton ton ⁻¹ biomass)	0.12
¹ Cost NaOH (\$ ton ⁻¹ biomass)	42
Dose Ca(OH) ₂ (ton ton ⁻¹ biomass)	0.11
² Cost Ca(OH) ₂ (\$ ton ⁻¹ biomass)	17
De-flocculation	
Dose HCl (ton ton ⁻¹ biomass)	0.05
³ Cost HCl (\$ ton ⁻¹ biomass)	40

590 ¹ NaOH industrial grade \$350 ton⁻¹591 ² Ca(OH)₂ slaked lime \$150 ton⁻¹592 ³ HCl industrial grade 35% \$250 ton⁻¹

593

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